

Population Dynamics and Harvest Management of the Continental Northern Pintail Population

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1. Executive Summary

Pursuant to requests from the Service Regulations Committee and the Pacific Flyway Council, we reviewed available information about northern pintail (*Anas acuta*) demography, population dynamics, and harvest. Based on this review, we suggest that several technical improvements in our ability to model pintail harvest dynamics be considered. In addition, we undertook an effort to evaluate pintail harvest potential based on these model improvements and to explore the impacts of these improvements on past and future pintail harvest management policy. Notable findings from this report, especially those that might warrant comment, are summarized below.

- *Breeding Population Survey Corrections.* There is general agreement among waterfowl scientists that the May breeding population survey undercounts pintails in dry years when pintails tend to settle farther north on the breeding grounds. We developed a method to correct the observed breeding population estimates for this bias. The effect of this correction is to remove some of the apparent sharp drops in pintail numbers during dry years. Further, this correction suggests that in recent years, there are 30-60% more pintails in the breeding population than the May surveys indicate.
- *Updated Recruitment, Harvest, and Population Models.* We developed improved methods to predict recruitment and harvest, and included these components in an updated population model for use in the pintail harvest strategy. The recruitment model uses latitude of the pintail population and the corrected breeding population size estimates as predictors. The harvest models identify a “season-within-a-season” effect in the Central and Mississippi flyways. The new population model predicts population change better than the previous model.
- *Pintail Harvest Potential.* Using the new pintail population model, we were able to analyze the harvest potential of the pintail population. There is evidence that the pintail population is settling, on average, about 2.4° of latitude farther north now than it did prior to 1975, possibly as a result of changes in habitat. This more northern distribution has resulted in lower reproduction, a 30-45% decrease in carrying capacity, and a 40-65% decrease in sustainable harvest potential.
- *Pintail Harvest Strategy.* If we embed these technical improvements into the current pintail harvest strategy, and take account of the post-1975 environmental conditions, we expect pintail season length to deviate from the AHM season length 13% of the time, the average observed pintail BPOP to be around 3.7 million, and the average annual continental harvest to be around 380,000. The frequency of pintail seasons-within-a-season would increase if the North American goal were removed from the AHM objective function.

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2. Introduction

North American northern pintails (*Anas acuta*) have shown a substantial population decline in the past 30 years, presumably due to anthropogenic changes to the landscape on the breeding grounds (Miller and Duncan 1999). Prior to this reduction in population size, pintails were an important component of the U.S. harvest, and there continues to be substantial interest in maintaining sustainable harvest of pintails.

In recognition of the poor status of pintails, in 1997 the USFWS developed a special harvest strategy (referred to as the “Interim Pintail Harvest Strategy”) that has remained in effect since then. The interim strategy underwent some minor technical changes in 2002 when the harvest models were updated. In the 2002-3 and 2003-4 hunting seasons, the Service set pintail regulations (“season-within-a-season”) that deviated from the strict prescriptions of the interim strategy, but remained true to the intent of the strategy. For the 2004-5 hunting season, the Service formally incorporated seasons-within-a-season as a component of the pintail strategy. In adopting those changes, the Service and others called for an analysis of the expected performance of the pintail strategy and consideration of technical modifications that could be made to improve it.

This report contains a thorough review of the demographic and harvest information available at the continental scale for northern pintails. In addition, it considers a number of technical modifications that might be made to the modeling framework used in the pintail harvest strategy. Finally, it begins to explore the harvest implications of the current pintail strategy, especially with regard to the impact of the long-term changes to the breeding habitat.

This report is currently undergoing formal peer-review through the US Geological Survey. We are releasing this draft prior to the completion of peer review so the Flyway technical committees have time to review it before their July 2005 meetings. We would welcome feedback from the technical committees, their individual members, or other interested parties. A final version of this report will be prepared in response to all comments received by 1 September 2005, including those that arise from the peer review process. We anticipate distribution of the final version of this report in fall 2005.

3. Estimates of Demographic and Harvest Parameters

To facilitate the development of technical improvements to the population modeling framework currently used in the Interim Pintail Harvest Strategy, we found it necessary to update several population parameter estimates based on contemporary information available from the banding and harvest survey programs. These analyses were conducted to extend the time series of demographic parameter estimates originally reported by Sheaffer et al. (1999), which included estimates of direct recovery rates, relative harvest vulnerabilities, age ratios, and annual survival and recovery rates. The primary purpose of this work was to update the population age ratio estimates for use in the development of a predictive model of pintail recruitment and to update our current knowledge of pintail annual survival and recovery rates.

Direct recovery rates and harvest vulnerabilities. We estimated direct recovery rates of pintails encountered in the United States during the hunting season from 1960 – 2003 (Appendix 2). Direct recovery rates were based on the number of pre-season bandings (July-September) and unsolicited recoveries of individuals shot or found dead during the hunting season (September – January). We calculated direct recovery rates (f) and corresponding variances in year t for each sex (s) and age class ($a = \text{young or adult}$) according to

$$f_{t,s,a} = m_{t,s,a} / R_{t,s,a} \quad \text{and} \quad (3.1)$$

$$Var(f_{t,s,a}) = f_{t,s,a}(1 - f_{t,s,a}) / R_{t,s,a} \quad (3.2)$$

where, $m_{t,s,a}$ = number of direct recoveries of individuals of sex s and age a in year t and $R_{t,s,a}$ = total number of birds of sex s and age a released in year t . We used direct recovery rate estimates of birds encountered in the US to calculate the relative vulnerability (young:adult) of pintails to the US harvest (Appendix 3). For each sex (s), we estimated the age-related vulnerability and corresponding variances in year t with

$$V_{t,s} = f_{t,s,young} / f_{t,s,adult} \quad \text{and} \quad (3.3)$$

$$Var(V_{t,s}) = \frac{Var(f_{t,s,young})}{f_{t,s,adult}^2} + \frac{f_{t,s,young}^2 Var(f_{t,s,adult})}{f_{t,s,adult}^4} \quad (3.4)$$

Population age ratios

We used the ratio of young to adult females in the US harvest to calculate pintail population age ratios ($R_{t,female}$) according to

$$R_{t,female} = A_{t,female} / V_{t,female} \quad (\text{Martin et al. 1979}), \quad (3.5)$$

where $A_{t,female}$ is the female age ratio in the US harvest in year t (Appendix 3).

Estimates of age-related female vulnerability to the US harvest have been variable, ranging from a low of 0.97 in 1982 to a high of 3.46 in 1992 (Figure 3.1 A). The female population age ratio from 1961 to 2003 is also variable (Fig. 3.1 B) with a long term mean equal to 1.01 (se = 0.061).

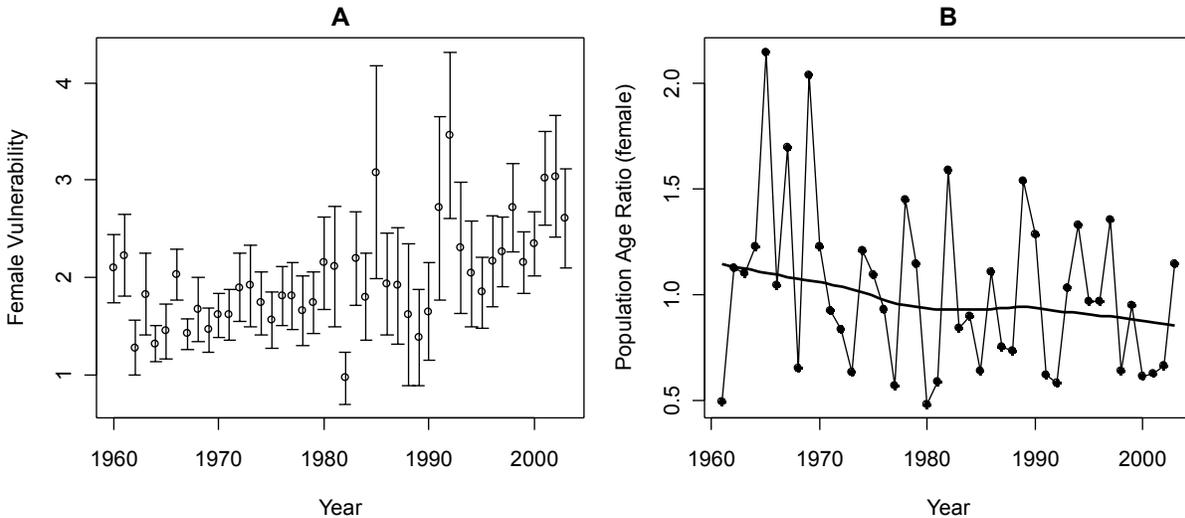


Figure 3.1. A. The relative vulnerability (young:adult) of female pintails to the US harvest from 1961 to 2003; error bars represent plus or minus one standard error. (B) The population age ratio (female) of pintails from 1961 to 2003 calculated from the US harvest age ratio and the female age-related vulnerability estimates. The solid line represents a loess smooth.

Annual survival and recovery rates

We calculated annual survival and recovery rates for each sex and age class using Brownie model H1 (Brownie et al. 1979) in program Mark (White and Burnham 1999). We used recoveries of individuals banded preseason in Canada and the United States that were shot or found dead during the hunting season.

Annual survival rates from 1960 to 2002 were highly dynamic across all sex and age classes (Figures 3.2 A1-4). In addition, the estimated confidence intervals were very large, making it hard to evaluate how annual survival of northern pintails has changed over time (Appendices 4-7). The large sampling errors may be attributed to the limitations of the pintail banding and recovery data. In contrast to the annual survival estimates, the recovery rate estimates suggest a common pattern of decreasing recovery rates from the 1970's through the 90's with large increases in recovery rates throughout the period of liberalization associated with the Adaptive Harvest Management program (Figures 3.2 B1-4). Our ability to determine if these changes are representative of actual changes in harvest rates or changes in reporting rates are constrained by the lack of reporting rate information. However, the recovery rate information does depict a common signal across all sex and age classes, including a consistent reduction during the restrictive seasons in 2002 and 2003. This evidence suggests that these estimates may be informative when evaluating how changes in pintail harvest policy may affect pintail recovery rates.

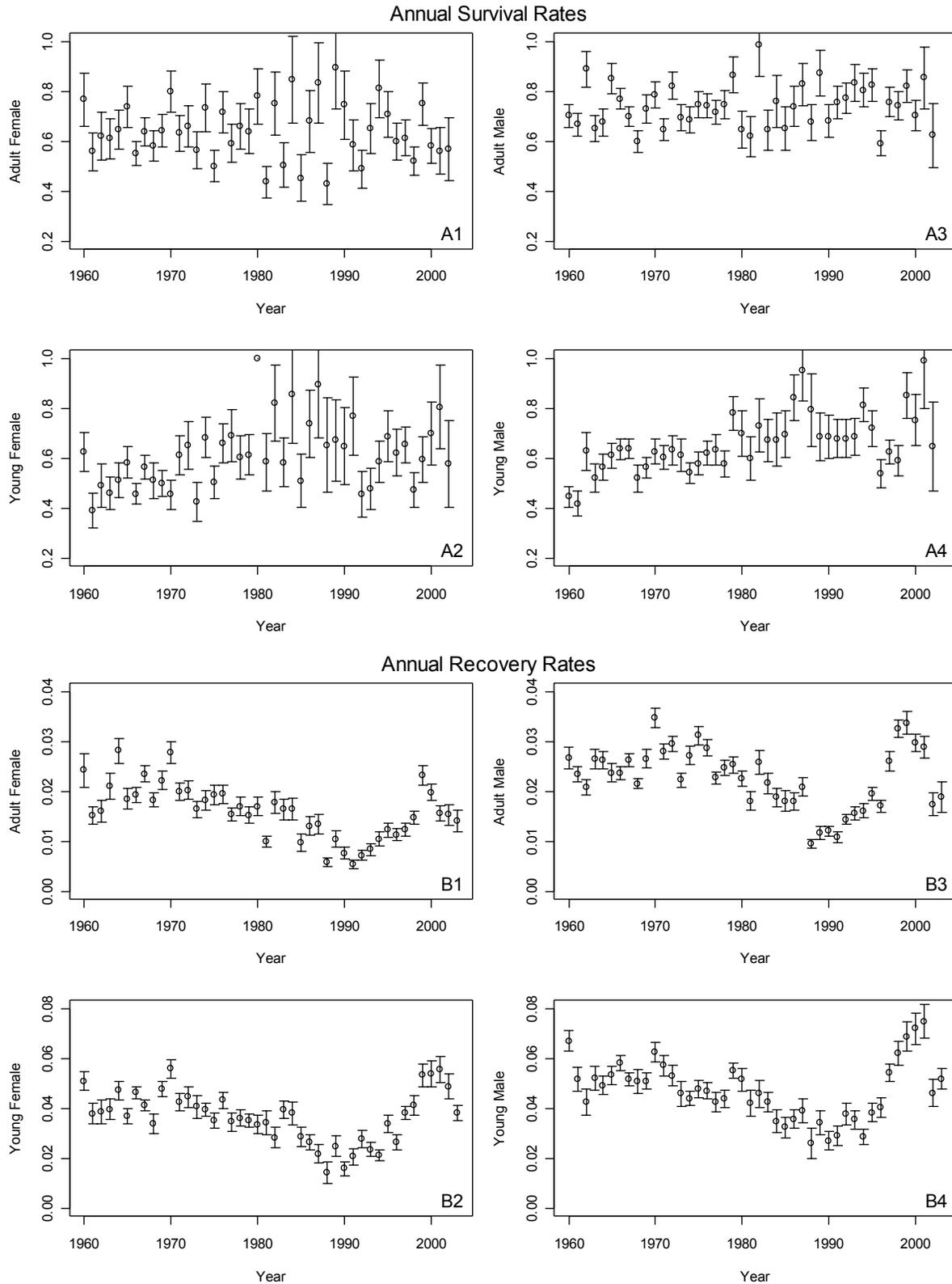


Figure 3.2. Annual survival (A1-4) and recovery rate (B1-4) estimates for adult and young northern pintails banded pre-season in the US and Canada and recovered during the hunting season from 1960-2003. Error bars represent plus or minus one standard error.

4. Harvest Models

We updated models to predict the total harvest of pintails in each Flyway as a function of Flyway-specific regulations. Our modeling approach was very similar to the analyses used to derive the current harvest models (Runge 2002), but included an additional 3 years of harvest data resulting from seasons specified under the Interim Pintail Harvest Strategy. The additional years of data (2001-2003) also provided experience with regulations that prescribed restrictive seasons (seasons-within-a-season). In addition, our updated analysis accounted for differences in harvest data collection methods associated with the change from the traditional Mail Questionnaire Survey (MQS) to the operational Harvest Information Program (HIP) survey conducted by the Branch of Harvest Surveys. We attempted to model the pintail harvest for each flyway based on a suite of predictors including season length, bag limit, May breeding population estimates, and the mid-winter survey. We used indicator variables to model the effect of a season-within-a-season (SIS) and to represent data collected with the HIP survey protocol. For each flyway-specific analysis, we were unable to detect a significant effect of the HIP survey relative to the MQS.

Central Flyway. Based on the previous analysis, we used data from 1985 – 2003 (Appendix 8) to avoid the complication of information collected under the Point System (10-bird bags). There is a significant season-within-a-season effect (SIS) effect ($P = 0.03$), when this variable is included in a model that predicts harvest as a function of the number of days and the bag limit (Table 4.1). We tested additional models that included interaction terms, but did not consider these models based on regression diagnostics.

Table 4.1. The results from fitting a linear model that predicts the pintail harvest in the Central Flyway as a function of season length (Days), bag limit (Bag) and a season-within-a-season effect (SIS).

```

Central Flyway: Harvest = Days + Bag + SIS
lm(formula = Harvest ~ Days + Bag + SIS, data = cf, subset = 7:28)
Residuals:
  Min       1Q   Median       3Q      Max
-28628  -3666  -1668    6135   21721

Coefficients:
              Estimate Std. Error t value Pr(>|t|)
(Intercept)   -95245      11776   -8.088 2.10e-07 ***
Days              2946         180  16.367 2.97e-12 ***
Bag             15228         3359   4.534 0.000257 ***
SIS             23136         9609   2.408 0.026996 *
---
Signif. codes:  '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

Residual standard error: 12010 on 18 degrees of freedom
Multiple R-Squared:  0.9437,    Adjusted R-squared:  0.9343
F-statistic: 100.6 on 3 and 18 DF,  p-value: 1.952e-11

Analysis of Variance Table
Response: Harvest
      Df    Sum Sq   Mean Sq  F value    Pr(>F)
Days    1 4.0128e+10  4.0128e+10  278.0835 2.164e-12 ***
Bag     1 2.5708e+09  2.5708e+09  17.8153 0.0005139 ***
SIS     1 8.3653e+08  8.3653e+08   5.7971 0.0269961 *
Residuals 18 2.5974e+09 1.4430e+08
AIC 481.3417
  
```

The resulting model to predict pintail harvest (H) in the Central Flyway is:

$$H = -95245.2 + 2946.285 \text{ Days} + 15228.03 \text{ Bag} + 23136.04 \text{ SIS} \quad (4.1)$$

where $Days$ is the season length, Bag is the bag limit and SIS is an indicator variable with a value equal to 0 (full season) or 1 (season-within-a-season). As a result of the significant SIS effect, the Central Flyway pintail harvest model predicts a larger harvest under a restrictive pintail season within an otherwise liberal duck season than the harvest predicted for a season that is restrictive for all ducks (Fig. 4.1).

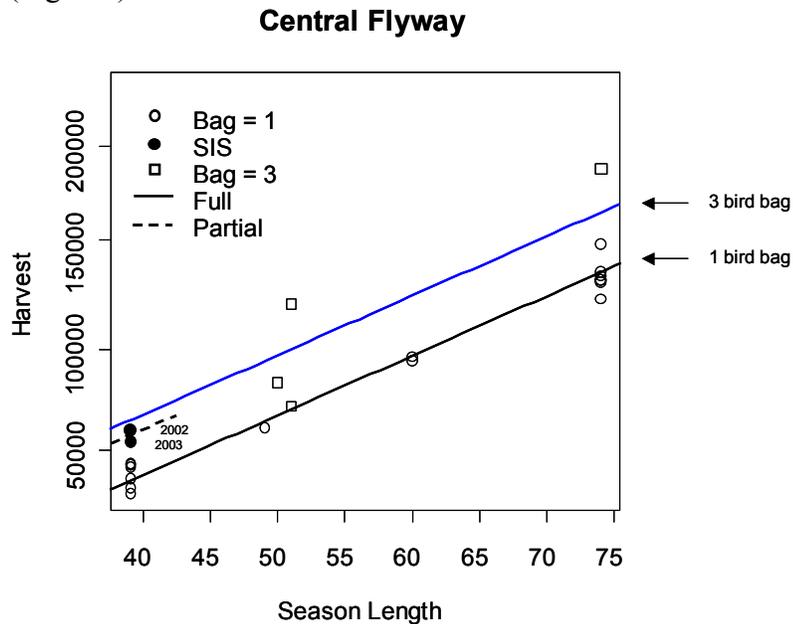


Fig. 4.1. Observed Central Flyway harvest (1985-2003) and predicted harvest calculated as a function of season length and bag limit of one (lower solid line) or three (upper solid line) based on the Central flyway harvest model, which includes a season within a season effect. Note, the harvest predicted under a season-within-a-season (dashed lined) assumes a 1-bird bag.

Atlantic Flyway. In 2002 and 2003 the Atlantic Flyway adopted a season-within-a-season, and subsequently experienced the largest recorded harvests associated with 30-day seasons and a 1-bird bag since 1979 (Appendix 9). As in the previous analysis (Runge 2002), an initial best subsets regression procedure with all predictor data failed to generate any reasonable models with season length or bag limit as predictors. As a result, we used methods from the previous analysis to reclassify bag limit information into 3 levels (1, 2, ≥ 3) using the entire data set (1979-2003). In addition, we tested additional models that included interaction terms, but did not consider these models based on regression diagnostics. The resulting model (Table 4.2) predicts harvest (H) as a function of season length ($Days$) and a linear effect of the reclassified bag limit ($BagClass$) values with

$$H = -2403.06 + 360.95 \text{ Days} + 5494 \text{ BagClass} . \quad (4.2)$$

Predictions from this model are shown in Fig. 4.2.

Table 4.2. The results from fitting a linear model that predicts the pintail harvest in the Atlantic Flyway as a function of season length (Days) and bag limit (BagClass) classified to 1, 2, or ≥ 3.

Atlantic Flyway: Harvest=Days+BagClass LINEAR
`lm(formula = Harvest ~ Days + BagClass, data = af, subset = c(1:28))`
Residuals:
 Min 1Q Median 3Q Max
-13490 -3991 -1167 3431 16336

Coefficients:
 Estimate Std. Error t value Pr(>|t|)
(Intercept) -2403.1 5219.3 -0.460 0.64919
Days 360.9 111.1 3.250 0.00329 **
BagClass 5494.0 1516.0 3.624 0.00129 **

Signif. codes: '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

Residual standard error: 6720 on 25 degrees of freedom
Multiple R-Squared: 0.555, Adjusted R-squared: 0.5194
F-statistic: 15.59 on 2 and 25 DF, p-value: 4.025e-05

Analysis of Variance Table
Response: Harvest

	Df	Sum Sq	Mean Sq	F value	Pr(>F)
Days	1	814734929	814734929	18.044	0.0002616 ***
BagClass	1	593024945	593024945	13.134	0.0012920 **
Residuals	25	1128828111	45153124		

AIC 577.8033

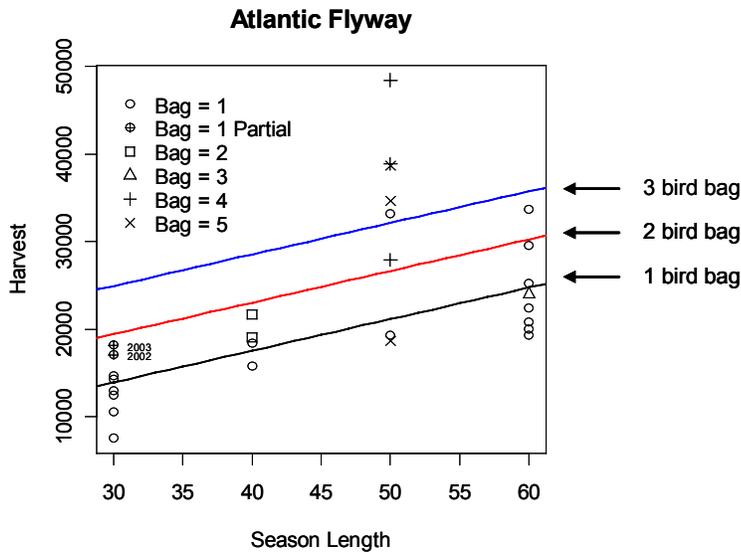


Fig. 4.2. Observed Atlantic Flyway pintail harvest (1979-2003) and predicted harvest calculated as a function of season length and a bag limit of one (lower line), two (middle line) or three (upper line) based on the Atlantic flyway harvest model.

Mississippi Flyway. Similar to the Atlantic Flyway, in 2002 and 2003 the Mississippi Flyway adopted a season-within-a-season, and subsequently experienced the largest recorded harvests associated with 30-day seasons and a 1-bird bag since 1979 (Appendix 10). As in the original analysis, we did not use information collected during the 10-bird bag limits under the Point

System (1979-1984). Regression diagnostics based on a model that included season length and bag limits as predictors indicated that the data collected in 1997 had a very high influence as measured by Cook's Distance. As a result, we dropped this data point and refit the models. The resulting model included a significant SIS effect ($P < 0.001$, Table 4.3).

Table 4.3. The results from fitting a linear model that predicts the pintail harvest in the Mississippi Flyway as a function of season length (Days), bag limit (Bag) and a partial season effect (SIS).

```
Mississippi Flyway: Harvest = Days + Bag + SIS
lm(formula = Harvest ~ Days + Bag + SIS, data = mf, subset = c(7:18, 20:28))

Residuals:
    Min       1Q   Median       3Q      Max
-31116 -10842   1402    8262  23728

Coefficients:
            Estimate Std. Error t value Pr(>|t|)
(Intercept) -59083.7   15867.2   -3.724  0.001689 **
Days         3413.5     284.8   11.986  1.02e-09 ***
Bag          7911.9     4966.9    1.593  0.129595
SIS          59510.1    12535.7    4.747  0.000187 ***
---
Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

Residual standard error: 15530 on 17 degrees of freedom
Multiple R-Squared:  0.8945,    Adjusted R-squared:  0.8759
F-statistic: 48.05 on 3 and 17 DF,  p-value: 1.626e-08

Analysis of Variance Table
Response: Harvest
      Df    Sum Sq   Mean Sq  F value    Pr(>F)
Days    1 2.9223e+10  2.9223e+10  121.1105  3.738e-09 ***
Bag     1 1.2393e+08  1.2393e+08   0.5136  0.4833040
SIS     1 5.4378e+09  5.4378e+09  22.5364  0.0001866 ***
Residuals 17 4.1020e+09  2.4129e+08
AIC 470.4898
```

The resulting model to predict pintail harvest (H) in the Mississippi Flyway is:

$$H = -59083.66 + 3413.49 \text{ Days} + 7911.95 \text{ Bag} + 59510.1 \text{ SIS} \quad (4.3)$$

where $Days$ is the season length, bag is the bag limit, and SIS is an indicator variable with a value equal to 0 (full season) or 1 (season-within-a-season). As a result of the significant SIS effect, the Mississippi Flyway pintail harvest model predicts a larger harvest under a restrictive pintail season within an otherwise liberal duck season than the harvest predicted for a season that is restrictive for all ducks (Fig. 4.3).

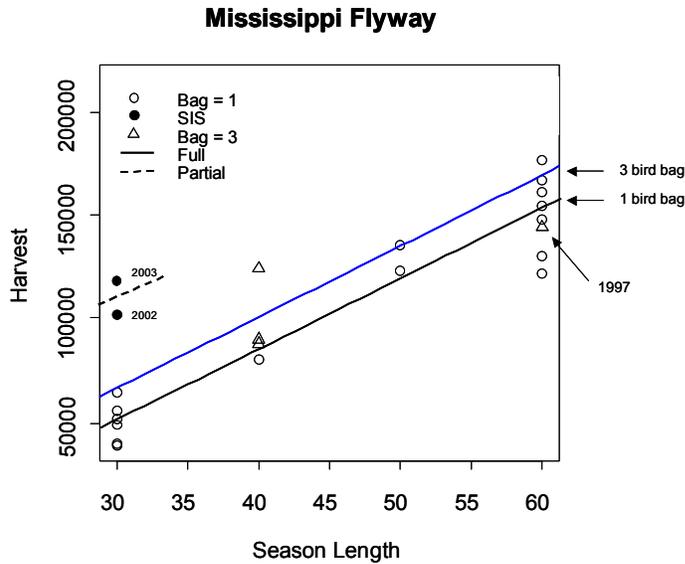


Fig. 4.3. Observed Mississippi Flyway harvest (1985-2003) and predicted harvest calculated as a function of season length and bag limit of one (lower solid line) or three (upper solid line) based on the Mississippi flyway harvest model, which includes a season within a season effect. Note, the harvest predicted under a partial season (dashed line) assumes a 1-bird bag.

Pacific Flyway. As in the original analysis (Runge 2002), we did not use information collected during the 7-bird bag limits in use from 1979-1983. Based on data from 1984-2003, there is only one year of experience with a 3-bird bag, two years with a 4-bird bag (both at the same season length), and two years with a 5-bird bag (but at different season lengths). The 5-bird bag data provides a little more information about the effect of season length at the higher bag limits. It is visually evident in Fig. 4.4 that there isn't a significant difference between 3, 4 or 5 in the bag, so we considered lumping those higher bag limits as we did in the Atlantic Flyway analysis. As a result, we reclassified bag limit information into 3 levels (1, 2, ≥ 3). In addition, we tested additional models that included interaction terms, but did not consider these models based on regression diagnostics. There was no evidence of a season-within-a-season effect ($P = 0.87$). The resulting model (Table 4.4) predicts harvest (H) as a function of season length ($Days$) and a linear effect of the reclassified bag limit ($BagClass$) values with

$$H = -12051.41 + 1160.96 Days + 73911.49 BagClass . \quad (4.4)$$

Predictions with this model are shown in Fig. 4.4.

Table 4.4. The results from fitting a linear model that predicts the pintail harvest in the Pacific Flyway as a function of season length (Days) and bag limit (BagClass) classified to 1, 2, or ≥ 3 .

Pacific Flyway: Harvest=Days+BagClass LINEAR

```
lm(formula = Harvest ~ Days + BagClass, data = pf, subset = 6:28)
```

Residuals:

```
      Min       1Q   Median       3Q      Max
-52775 -11367   2798   9579  51194
```

Coefficients:

```
              Estimate Std. Error t value Pr(>|t|)
(Intercept) -12051.4    22479.9  -0.536 0.597803
Days          1161.0     254.8    4.556 0.000192 ***
BagClass      73911.5    6313.7   11.707 2.10e-10 ***
```

```
Signif. codes:  0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
```

Residual standard error: 24740 on 20 degrees of freedom

Multiple R-Squared: 0.8994, Adjusted R-squared: 0.8893

F-statistic: 89.4 on 2 and 20 DF, p-value: 1.062e-10

Analysis of Variance Table

Response: Harvest

	Df	Sum Sq	Mean Sq	F value	Pr(>F)	
Days	1	2.5566e+10	2.5566e+10	41.767	2.659e-06	***
BagClass	1	8.3887e+10	8.3887e+10	137.043	2.103e-10	***
Residuals	20	1.2242e+10	6.1212e+08			

AIC 535.4028

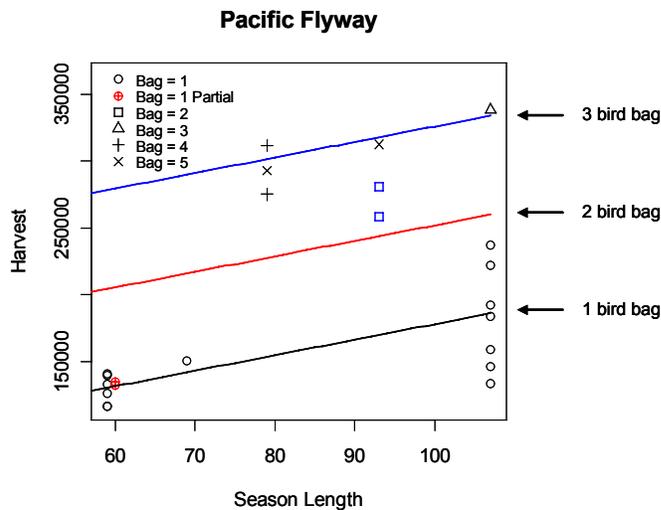


Fig. 4.4. Observed Pacific Flyway pintail harvest (1984-2003) and predicted harvest calculated as a function of season length and a bag limit of one (lower line), two (middle line) or three (upper line) based on the Pacific flyway harvest model.

Total Harvest Predictions. The updated harvest models provide a predictive relationship between flyway-specific hunting regulations and the expected harvest of pintails. The population model used under the current Interim Pintail Harvest Strategy assumes that the annual harvest from Alaska and Canada is a fixed value equal to 67,000 birds, based on modifications

made in February 2002 (for comparison, the average harvest in Alaska and Canada from 1996 through 2003 was 65,910, see Appendix 12). We used this fixed harvest amount (67,000) for Alaska and Canada and the updated harvest models to predict the total continental pintail harvest expected under a range of different regulatory options (Table 4.5). These results indicate a larger expected harvest under a season-within-a-season as compared to the expected harvest under an overall restrictive framework for all duck species. Caution is also warranted when evaluating the predictions under a liberal season with a two-bird bag because we have very little experience with these regulations.

Table 4.5. Predicted harvest of pintails calculated with the updated Flyway-specific harvest models, as a function of different regulatory options.

Flyway	Harvest Regulation ¹				
	R1	R1SIS	L1	L2	L3
Central	34,877	58,013	137,987	153,215	168,443
Atlantic	13,918	13,918	24,745	30,239	35,733
Mississippi	51,233	110,743	153,638	161,550	169,462
Pacific	131,520	131,520	186,087	259,999	333,910
Total US	231,548	314,194	502,457	605,003	707,548
Alaska/Canada	67,000	67,000	67,000	67,000	67,000
Total	298,548	381,194	569,457	672,003	774,548

¹ R1 = Restrictive season with a one bird bag.

R1SIS = Restrictive season within an overall liberal framework (season-within-a-season) with a 1 bird bag.

L1 = Liberal season with a 1-bird bag.

L2 = Liberal season with a 2-bird bag.

L3 = Liberal season with a 3-bird bag.

5. Accounting for Overflight Bias

One of the major concerns with the current pintail harvest policy is that it relies on population size estimates from the May aerial survey that are biased. In dry years on the prairie when pintails settle farther north, the BPOP tends to drop at a rate that cannot be explained by population dynamics alone. There is a notable negative relationship between BPOP estimates and latitude (Fig. 5.1a), suggesting that BPOP is lower when latitude is higher. This relationship could be explained by at least two hypotheses: (i) the overflight hypothesis—when pintails settle farther north, a smaller proportion are counted, thus the population estimate is lower; or (ii) the recruitment hypothesis—when pintails settle farther north, the continental production is lower, and because dry periods tend to last several years, this eventually results in lower population sizes, thus inducing the negative correlation observed. We can distinguish these two hypotheses by detrending the BPOP and latitude time series, that is, by looking at the one-year changes in BPOP and latitude against each other (Fig. 5.1b). Since the recruitment hypothesis would produce the negative relationship in Fig. 5.1a through a time-delayed mechanism, if that hypothesis holds, the detrended series should not exhibit the negative correlation. The detrended series (Fig. 5.1b) show that all other things being equal, when the latitude increases between year $t-1$ to year t , the BPOP estimate drops, supporting the overflight hypothesis.

The relationship between the detrended BPOP and latitude series suggests a method to correct the BPOP estimates for this overflight bias. The line fit through the detrended series has a significant slope of -0.0741 ; for every 1 degree increase in latitude, the logarithm of the BPOP estimate drops by 0.0741 (this is equivalent to a 7.1% decrease). Mathematically,

$$\Delta \ln B_t = -0.0741 \Delta L_t \quad (5.1)$$

where B_t is the breeding population size estimate in year t , and L_t is the latitude of the center of the distribution of the breeding population in year t . Equation 5.1 can be expanded to

$$\ln B_t - \ln B_{t-1} = -0.0741(L_t - L_{t-1}). \quad (5.2)$$

This equation (5.2) could be used to correct a BPOP in one year to put it on the same scale as the BPOP in the previous year (by “same scale”, we mean relative to the latitude). By extension, we could modify this equation to express this year’s BPOP relative to some other scale. We’d like to express it on the “true” scale, but there is no point that we know is completely unbiased in estimating the pintail population size. As an approximation, we chose 1969, the year in which the latitude of the pintail population was as far south as it has ever been observed (51.68°N). Thus, we can modify equation 5.2 to

$$\begin{aligned} \ln B_{obs} - \ln B_{true} &= -0.0741(L_{obs} - L_{true}) \\ \ln B_{obs} - \ln B_{true} &= -0.0741(L_{obs} - 51.68) \end{aligned} \quad (5.3)$$

and then a little algebra allows us to develop an expression for the corrected BPOP estimate:

$$\begin{aligned} \ln B_{true} &= \ln B_t + 0.0741(L_{obs} - 51.68) \\ B_{true} &= e^{\ln B_t + 0.0741(L_{obs} - 51.68)} \end{aligned} \quad (5.4)$$

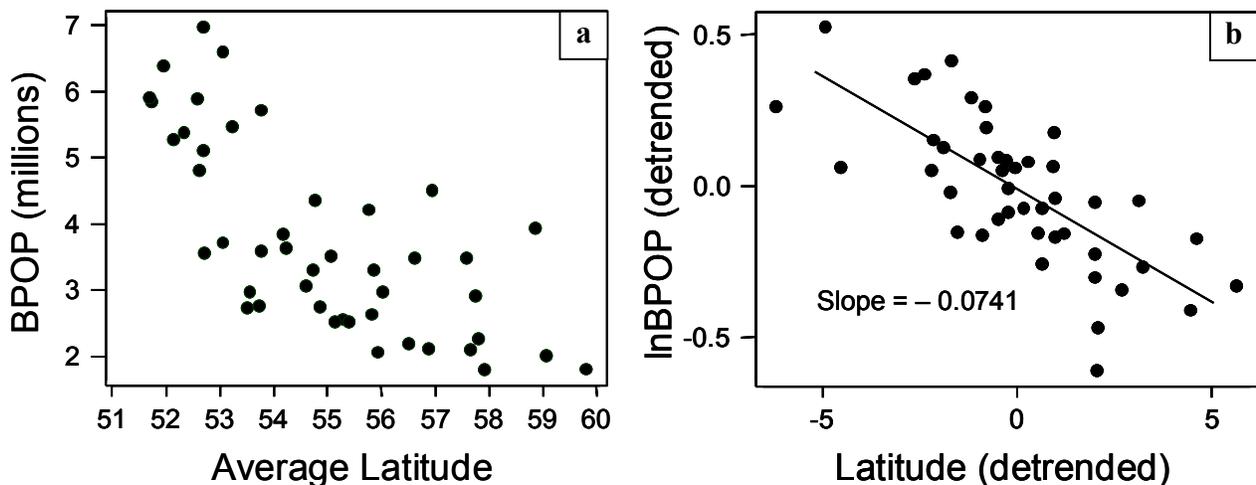


Fig. 5.1. Observed pintail breeding population size plotted against the latitude of the center of the distribution of the pintail population, 1960-2004. (a) Observed values. (b) Detrended values. The logarithm of the ratio of the breeding population size in year t to the population size in year $t-1$ is plotted against the difference in latitude between year t and year $t-1$. The fitted line has a slope of -0.0741 .

The problems of overflight bias can be seen in the graph of observed breeding population size estimates against time (Fig. 5.2, solid line): there are years (e.g., 1968, 1973, and 1977) when a large portion of the pintail population seems to disappear, only to reappear a year later; and the decline between 1975 and 1990 was accompanied by an increase in latitude, raising the question of whether the apparent decline was exaggerated by an increasing trend in overflight. The corrected estimates of the breeding population size (Fig. 5.2, dotted line) largely remove these patterns. In recent years, the correction suggests there are 30-60% more pintails in the breeding population than the May surveys indicate (see corrected values in Appendix 1).

This correction to the aerial survey dovetails with the findings from the satellite telemetry project of Miller et al. (2003), who state that the “May Survey did not efficiently account for pintails.” One hypothesis for the overflight bias is a detectability issue in the northern strata. In the prairie strata, the May aerial survey, with coordinated annual ground counts, estimates the pintail population relatively accurately. But in the northern strata, perhaps the aerial survey estimates miss a substantial portion of the population. Potential reasons for this include: the visibility adjustment factors are not estimated regularly and might be biased; the strata may not be located in all the areas pintails use; the timing of surveys may not be optimized for pintail migration; and the area expansion factors for some of the northern strata may be too small. Whatever the reason, if the estimates from the southern strata are largely unbiased and the estimates from the northern strata are negatively biased, then the degree of overall bias depends on the distribution of the pintail population among northern and southern strata, a distribution that is captured to a large degree by the latitude of the breeding population. This hypothesis underlies our choice of 1969 as the reference point in the correction; because that was the year with the southernmost observed pintail distribution, the correction in all other years produces estimates that are higher than the observed survey estimates.

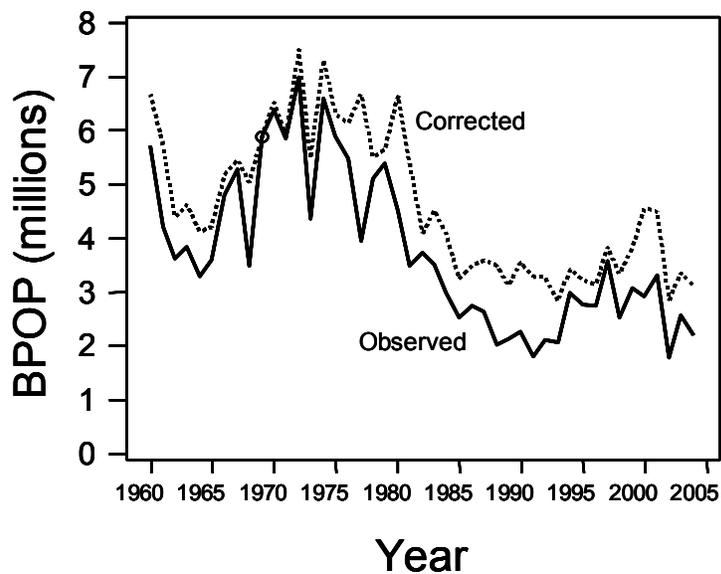


Fig. 5.2. Observed and corrected pintail breeding population sizes, 1960-2004. The correction is based on the average latitude of the breeding population, and is calibrated so the two curves coincide in 1969 (the year with the most southern observed latitude).

The derived correction (equation 5.4) is an attempt to empirically remove this overflight bias. Use of this correction is supported in several ways: (i) the correction is based on a statistical relationship that supports the overflight hypothesis; (ii) the corrected BPOP estimates are intuitively appealing, because to a large degree they remove the disturbing patterns in the time series of observed BPOP estimates; and (iii) use of these corrected BPOPs in a pintail population model significantly improves our ability to predict changes in the population size (see section 8.1, “Predictive Ability of Population Models”, below). Use of this correction could be questioned for at least two reasons: it involves an arbitrary choice of reference point; and it is an empirical, rather than mechanistic, correction. We believe it is important for the US Fish and Wildlife Service and the Flyways to evaluate this correction and determine whether the corrected breeding population size estimates should be the basis of decision-making, especially in relation to harvest management.

6. Recruitment Models

The recruitment model currently used in the pintail harvest strategy was developed in 1997. There are three important reasons to consider updating it at this time. First, there is additional data (7 years worth) to include in a new analysis. Second, we now recognize that the current recruitment model is not explicitly density-dependent, but is implicitly so because one of the predictors in the model is strongly correlated with pintail BPOP. Third, with the development of the overflight-bias correction, we have a new predictor to consider including.

The recruitment model currently in use predicts the vulnerability-adjusted age-ratio (the fall age-ratio) as a function of the latitude of the center of the breeding population distribution, the variance of that mean latitude, and the ratio of the population size in northern vs. southern strata. The current model is

$$\ln R_t = 7.4076 - 0.1372L_t + 0.033P_{32}Var(L_t) \quad (6.1)$$

where R_t is the female population age-ratio (see section 3 above), L_t is the latitude of the center of the pintail distribution, and P_{32} is the ratio of pintails in northern to southern strata. The variance of the latitude is not a measure of the latitudinal spread of the pintail population, but a measure of sampling error in estimating L_t ; thus, it is inversely proportional to the sample size, that is, the pintail population size. In this way, the variance term is actually density-dependent, and induces density-dependence in the recruitment model. For the purposes of analysis, however, it is much easier to have the density-dependence be explicit, rather than hidden in one of the other variables.

We used the analysis of Sheaffer et al. (1999) as the basis for developing a new recruitment model. After extensive consideration of many predictors, they found that latitude and pintail BPOP were the most important predictors of recruitment, with some additional explanatory power provided by a few other environmental variables. We substituted the latitude-corrected pintail BPOP (see section 5 above) for the observed pintail BPOP, after analysis showed that it was a better predictor. The modified recruitment model is

$$\ln R_t = 7.605 - 0.1318L_t - 0.0921B_t^{adj} \quad (6.2)$$

where B_t^{adj} is the latitude-adjusted breeding population size (from equation 5.4). Both the latitude and the BPOP effects are highly significant ($P < 0.001$ and $P = 0.014$, respectively). This model explains 48.5% of the variation in the observed recruitment data (1961-2003), approximately the same amount explained by the model currently in use (equation 6.1). That is, in any given year, the predictions from these two models are very similar, but the new model has a clear mechanistic explanation and is more amenable to the types of analyses we commonly employ with duck population models. Regression diagnostics did not reveal any violations of the assumptions of the linear model. In particular, it's important to note there was no evidence that the effects of latitude or population size on recruitment had changed over time.

To explore uncertainty in this recruitment relationship, we considered parameter values on the 80% confidence ellipsoid, found using

$$(\beta - \hat{\mathbf{b}})' \mathbf{X}'\mathbf{X}(\beta - \hat{\mathbf{b}}) \leq ps^2 F_{p, n-p, 0.80} \quad (6.3)$$

where β is the vector of parameter estimates on the 80% confidence ellipsoid, $\hat{\mathbf{b}}$ is the vector of maximum likelihood estimates for the parameters, \mathbf{X} is the design matrix for the linear model, p is the number of parameters, s^2 is the mean squared error from the regression, and F refers the the F -statistic with p and $n-p$ degrees of freedom (Draper and Smith 1981). Five alternative models were considered (Table 6.1); their implications for the estimation of carrying capacity are discussed later in this report. There are an infinite number of parameter combinations on the 80% confidence ellipsoid; we chose these 4 (models 1-4) to emphasize uncertainty in the latitude effect.

Table 6.1. Alternative recruitment models used to capture parametric uncertainty in the effects of density and latitude.

Model	Intercept	Latitude slope	Density Slope	Location
0	7.6048	-0.1318	-0.0921	MLE
1	5.4693	-0.0919	-0.1052	80% CE
2	9.7407	-0.1718	-0.0791	80% CE
3	4.8636	-0.0845	-0.0615	80% CE
4	10.3455	-0.1792	-0.1227	80% CE

7. Evidence of Landscape Changes

In the last decade or so, there has been considerable attention focused on understanding the historic decline in pintail numbers and the failure of the pintail population to recover in the late 1990s when record water levels occurred in the Canadian prairies. There is emerging consensus among pintail biologists that the problem is localized to the breeding grounds (Miller and Duncan 1999, Miller et al. 2003). Since the 1950s, there has been a steady conversion of grasslands to cultivation in the western Canadian prairie pothole region, a trend that has reduced the availability of secure upland nesting habitat. The exact demographic mechanism that has driven the change in pintail dynamics remains open for discussion, but several strong hypotheses

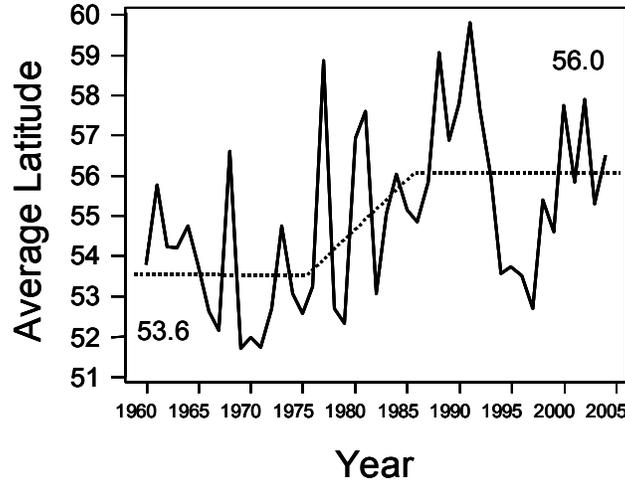


Fig. 7.1. Latitude of the center of the pintail distribution, 1960-2004. There is evidence that the average latitude shifted northward by 2.4 degrees between 1975 and 1985.

have been advanced, most of which involve long-term reductions in recruitment brought about by landscape change.

There is suggestive evidence on the continental scale that pintails have exhibited a behavioral response to the landscape changes. The distribution of the pintail breeding population appears to have shifted northward by about 2.4° latitude. Prior to 1975, the average latitude of the center of the pintail breeding distribution (as measured by the May surveys) was 53.6°N (with a standard deviation of 1.49 and non-significant autocorrelation of $\rho = 0.01$). Since 1985, that average has been 56.0°N with an increased variance ($SD = 1.93$) and evidence of stronger autocorrelation ($\rho = 0.51$, $P = 0.02$) (Fig. 7.1). In 1996-7, when water returned to the prairies, pintails did distribute farther south, but not as far south as in the early 1970s. Possibly, pintails are responding to the landscape changes by choosing to overfly the prairies to a greater degree than in the past, even under similar water conditions. An alternate hypothesis is that the component of the population that has the highest fidelity to prairie nesting areas has decreased, while the northern breeding component has remained stable; the net effect is a northward shift in average latitude.

Whatever the reason for the shift in latitude, the implication relative to recruitment is the same: recruitment in the average year, under similar densities and water conditions, has decreased. The best recruitment model (see section 6 “Recruitment Models”) shows a negative effect of latitude: when the latitude in a particular year increases, recruitment decreases, all other things being equal. Thus, this increase in average latitude implies a decrease in average recruitment. The ramifications of this shift on population dynamics and harvest potential are explored below (section 8.2 “Harvest Yield Analysis”).

The preceding argument implies a change in the relationship between pintail distribution and prairie water. The number of ponds in Canada counted during the May aerial surveys is often used as a measure of water on the prairie breeding grounds, and is a central state variable in the determination of harvest regulations for mid-continent mallards. While pintails are often associated with more ephemeral water bodies than those measured by “Canadian Ponds,” there is nevertheless an indication that pintails are settling at higher latitudes relative to the number of

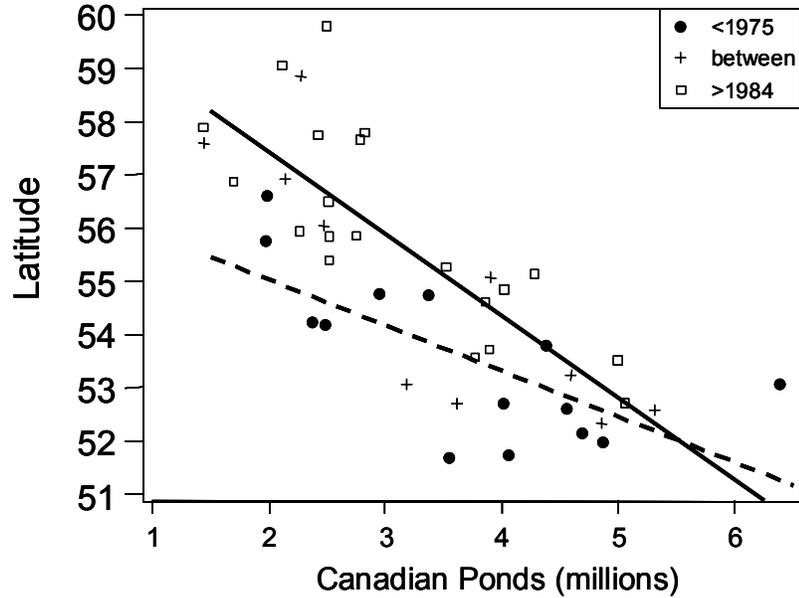


Fig 7.2. Latitude of the center of the pintail distribution against Canadian Ponds, 1961-2004. The fitted regression lines correspond to the periods 1961-1974 (dashed line) and 1975-2004 (solid line).

ponds than they were in the past (Fig. 7.2). We explored a large number of linear models fit to these data; the best fit was obtained using a model that distinguished the time period before 1975 from the time period including and after 1975. Prior to 1975, pintail distribution (as measured by average latitude of the breeding population) could be predicted from Canadian Ponds with

$$L_t = 56.7287 - 0.8567P_t \tag{7.1}$$

and the distribution after 1974 could be predicted with

$$L_t = 60.5474 - 1.5516P_t \tag{7.2}$$

where L_t is the latitude of the pintail population in year t and P_t is the number of Canadian Ponds. For both equations, the prediction variance is $(1.512)^2$.

There is not evidence that the mean number of Canadian ponds differed before and after 1975 (Table 7.1). Further, as noted in section 6, there is not evidence that the effect of pintail latitude on recruitment differed before and after 1975 (i.e., there is no latitude*time interaction in the regression in equation 6.2). Thus, the reduction in recruitment appears to be driven by a shift in the average latitude.

Table 7.1. Descriptive statistics for Canadian Ponds and pintail latitude, 1961-2004. The mean, standard deviation, and first-order autoregressive coefficient (ρ) are shown for three time periods.

	Canadian Ponds			Latitude		
	Mean	Std	ρ	Mean	Std	ρ
1961-1974	3.688 †	1.267	0.2044	53.569 ‡	1.549	-0.0063
1975-2004	3.186 †	1.112	0.2478	55.605 ‡	2.124	0.2683
All years	3.346	1.172	0.3182	54.957	2.165	0.3915

† Mean ponds do not differ significantly between these time periods ($P = 0.22$).

‡ Mean latitude *does* differ significantly between these time periods ($P = 0.001$).

8. Population Model

The operational population model that has formed the basis for the Interim Pintail Harvest Strategy is

$$\hat{B}_{t+1}^{obs} = \left[s_s B_t^{obs} (1 + \gamma_R \hat{R}_t^{old}) - \hat{H}_t^{old} / (1 - c) \right] s_w \quad (8.1)$$

where B_t^{obs} is the observed pintail breeding population size in year t , s_s and s_w are the summer and winter survival rates, respectively, γ_R is a bias-correction constant for the age-ratio, c is the crippling loss rate, and \hat{R}_t and \hat{H}_t are the predicted age-ratio and continental harvest based on the “old” models. Discussion of the recruitment and harvest models is found in preceding sections. The population model in the interim pintail harvest strategy uses the following constants: $s_s = 0.7$, $s_w = 0.93$, $\gamma_R = 0.8$, and $c = 0.2$.

We made three changes to this population model to incorporate analyses detailed in earlier sections: (1) we based the model on the latitude-corrected population sizes (equation 5.4) rather than the observed population sizes; (2) we used the new recruitment model (equation 6.2), which is explicitly density-dependent; and (3) we used the new harvest models (equations 4.1-4.4). Thus, the modified population model is

$$\hat{B}_{t+1}^{adj} = \left[s_s B_t^{adj} (1 + \gamma_R \hat{R}_t^{new}) - \hat{H}_t^{new} / (1 - c) \right] s_w \quad (8.2)$$

where B_t^{adj} is the latitude-adjusted pintail breeding population size in year t , \hat{R}_t and \hat{H}_t are based on the new models, and all the constants have the same values as previously. Expanding this equation to incorporate the new recruitment model,

$$\hat{B}_{t+1}^{adj} = \left[s_s B_t^{adj} \left(1 + \gamma_R e^{b_0 - b_1 B_t^{adj} - b_2 L_t} \right) - \hat{H}_t^{new} / (1 - c) \right] s_w \quad (8.3)$$

it becomes clear that this population model is explicitly density-dependent, and contains a term (the latitude of the breeding population, L_t) that governs environmental effects on the population dynamics.

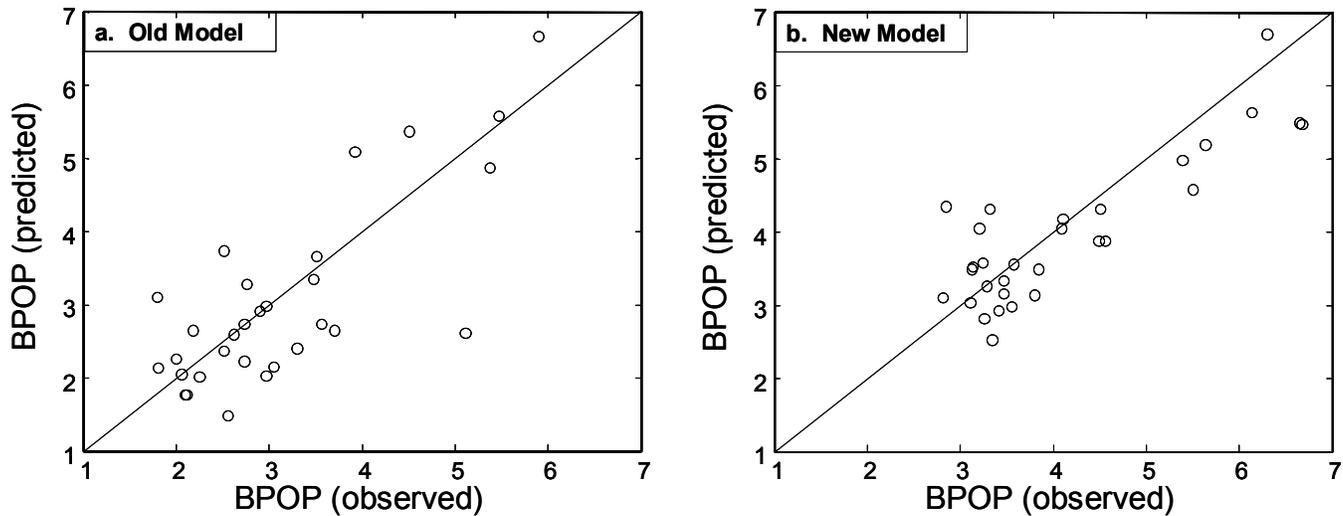


Fig. 8.1. Predictive ability of old and new pintail population models, 1974-2004. (a) Old model. Observed and predicted values on the *observed* BPOP scale, in millions. (b) New model. Values on the *corrected* BPOP scale.

8.1. Predictive Ability of Population Models

We were interested in evaluating the performance of the updated population model compared to the current model used in the Interim Pintail Harvest Strategy, and to other candidate predictors of pintail population size. To evaluate the predictive ability of the original and modified population models, we compared the predictions of the next year’s BPOP from each model to the observed BPOP values (Fig. 8.1). In a perfect model, the predictions will be equal to the observed values (that is, they’ll fall on the $y = x$ line in the graphs shown). Both models appear to be unbiased in their predictions, because there is no evident directional departure from the expected values. The predictions from the original (“old”) model show greater variance from the observed values than the predictions from the modified (“new”) model, indicating that the modified model is a more precise predictor than the old model.

To quantify the differences in predictive ability, we calculated the sums of squared error between the predictions and the observed values. In order to make a direct comparison between the two models, the predictions needed to be made on the same scale, so we converted the corrected BPOP estimates from the new model to the original scale, using the inverse of the overflight correction (the inverse of equation 5.4). We also considered a number of other predictors for next year’s BPOP (B_{t+1}), including the observed BPOP in year t (model 2) and the observed BPOP in year t corrected for the change in latitude between t and $t+1$ (model 3). For the old and new models, we considered the predictions using the recruitment model, and predictions using the observed age-ratios instead. We compared all the sums of squared error to the “total sum of squares,” which is the same as thinking of the mean BPOP as the predictor (model 1), and calculated a metric which is like R^2 , in that it measures the amount of variation explained by the model (Table 8.1).

The most important information for predicting BPOP ($t+1$) is the current year’s BPOP—model 2 explains about 63% of the variation. Using the overflight-correction to put BPOP(t) and BPOP($t+1$) on the same scale increases R^2 to 78% (model 3). Because the old model (model 4)

does not contain an overflight correction, it performs worse than the overflight-corrected BPOP(t), with an R^2 of only 48%; in fact, this is worse than simply using the observed BPOP from year t . The new model, back-transformed to the observed BPOP scale (model 6), was the best predictor, and explained 83% of the variation in BPOP($t+1$). Interestingly, the old and new models both performed better when using their models of recruitment, rather than the actual observed values for recruitment; we have not yet determined why this is the case.

In summary, when the technical modifications are included in the population model, the predictive ability of the model increases greatly (from R^2 of 48% to 83%). The vast portion of this improvement is due to use of the overflight-correction.

Table 8.1. Precision of various predictors of pintail breeding population size. The “ R^2 ” value reflects the proportion of the observed variation captured by the particular predictor, and is calculated as: $1 - \text{SSE}(\text{model } x) / \text{SSE}(\text{model } 1)$.

Predictive Model for observed BPOP ($t+1$)	SSE	“ R^2 ”
(1) Mean observed BPOP	37.09	0.0%
(2) Observed BPOP (t)	13.82	62.7%
(3) Observed BPOP (t), adjusted for Lat ($t+1$)	8.15	78.0%
(4) Old Model, using predicted R	19.19	48.3%
(5) Old Model, using observed R	25.65	30.8%
(6) New Model, using predicted R	6.46	82.6%
(7) New Model, using observed R	10.62	71.4%

8.2. Harvest Yield Analysis

Because the modified population model (equation 8.3) is explicitly density-dependent, its equilibrium harvest dynamics can be readily calculated (Runge and Johnson 2002). We calculated the carrying capacity by finding the equilibrium population size in the absence of harvest. Because the recruitment is a function of the latitude of the breeding population, carrying

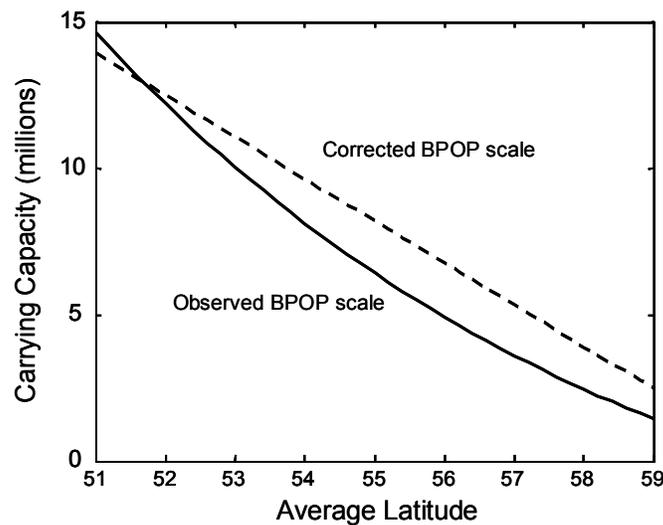


Fig. 8.2. Equilibrium population size in the absence of harvest (carrying capacity) as a function of average latitude of the breeding population.

capacity is a function of average latitude (Fig. 8.2). As the average latitude of the pintail distribution moves northward, the continental carrying capacity decreases. With an average latitude of 53.6 (as observed prior to 1975), the carrying capacity is 10.2 million on the latitude-corrected scale (8.9 million on the observed BPOP scale). With an average latitude of 55.6 (as observed since 1975), the carrying capacity is 7.3 million on the latitude-corrected scale (5.5 million on the observed scale). That is, if changes in the landscape have produced the shift in pintail distribution discussed above (section 7), then the carrying capacity (as measured by the equilibrium population size in the absence of harvest) has decreased by 28.5%. On the observed BPOP scale, the overflight bias accentuates this loss of carrying capacity, making it appear to be a 38.5% loss.

Loss of carrying capacity has direct implications for the productive capacity of the population and hence for the harvestable surplus. We investigated these implications by deriving harvest yield curves from the population model. To do this, we varied the harvest rate and calculated the corresponding equilibrium population size and sustainable annual harvest (Fig. 8.3). In each of the four curves in Fig. 8.3, the rightmost point shows the equilibrium population size in the absence of harvest (the carrying capacity). As harvest rate increases, the curve moves up and to the left, that is, the equilibrium population size decreases and the sustainable annual harvest increases. At some point, the maximum sustainable harvest is reached; beyond this point, increases in harvest rate continue to lower the equilibrium population size, but now, sustainable harvest also decreases. The two upper curves show the sustainable annual harvest as a function of the corresponding equilibrium population size, when the average latitude of the BPOP is 53.6°N (the average prior to 1975): the carrying capacity (equilibrium population size in the absence of harvest) was about 10.2 million (on the corrected scale, dotted line; 8.9 million on the observed scale), the maximum yield occurred when the population was held around 4.4 million (3.8 million on the observed scale), and the maximum sustainable annual harvest was more than what we would expect to achieve with a liberal season and three-bird bag in all Flyways.

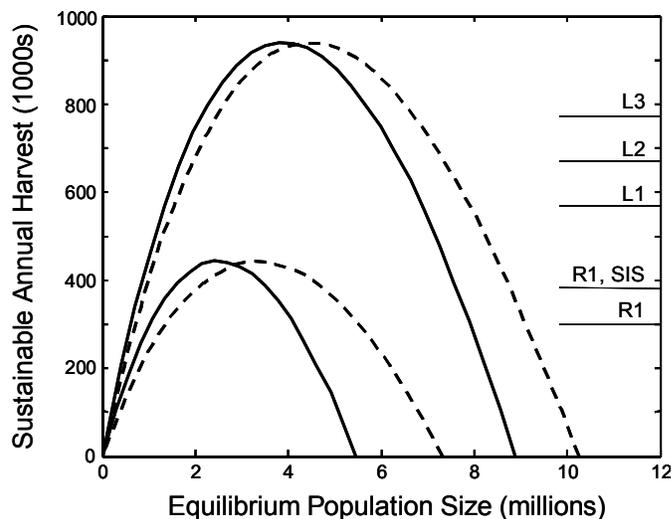


Fig. 8.3. Sustainable annual harvest of pintails as a function of equilibrium population size (solid lines: observed BPOP scale; dashed lines: latitude-corrected BPOP scale). The top curves are under environmental conditions that produce an average latitude of 53.6; the bottom curves assumes an average latitude of 55.6. The five horizontal reference lines show the expected continental harvest under a series of harvest packages (see Table 4.5), including a restrictive pintail season within a liberal duck season (“R1, SIS”).

The two lower curves in Fig. 8.3 correspond to an average latitude of 55.6°N (the average since 1975). The carrying capacity has been reduced to about 7.3 million (on the corrected scale, 5.5 million on the observed scale), the maximum yield occurs when the population is held at 3.3 million (2.4 million on the observed scale), and the maximum sustainable harvest is a little bit more than what we expect under a restrictive pintail season within a liberal duck season. These graphs show the central tendencies under average environmental conditions; in wet years, of course, the harvest potential could allow for a liberal season. We think this graphical depiction of the change in harvest potential as a result of a long-term change in pintail dynamics has immediate harvest management implications and is important to keep in mind as we continue to evaluate pintail harvest policies.

Uncertainty about the harvest yield curves in Fig. 8.3 derives from several sources: uncertainty about the model structure itself; uncertainty about the parameter estimates within the model; and uncertainty about the appropriate mean latitude to consider. Two of the most important components of uncertainty in this model are the effects of density and latitude on recruitment. Density-dependence in recruitment is a key determinant of the yield curve, and the latitude effect is the key determinant of the difference between the yield curves pre- and post-1975. To understand the effects of parametric uncertainty in the recruitment relationship on the yield curves, we considered 4 additional recruitment models (Table 6.1), substituted them into the population model, and calculated the corresponding yield curves (Fig. 8.4). The additional

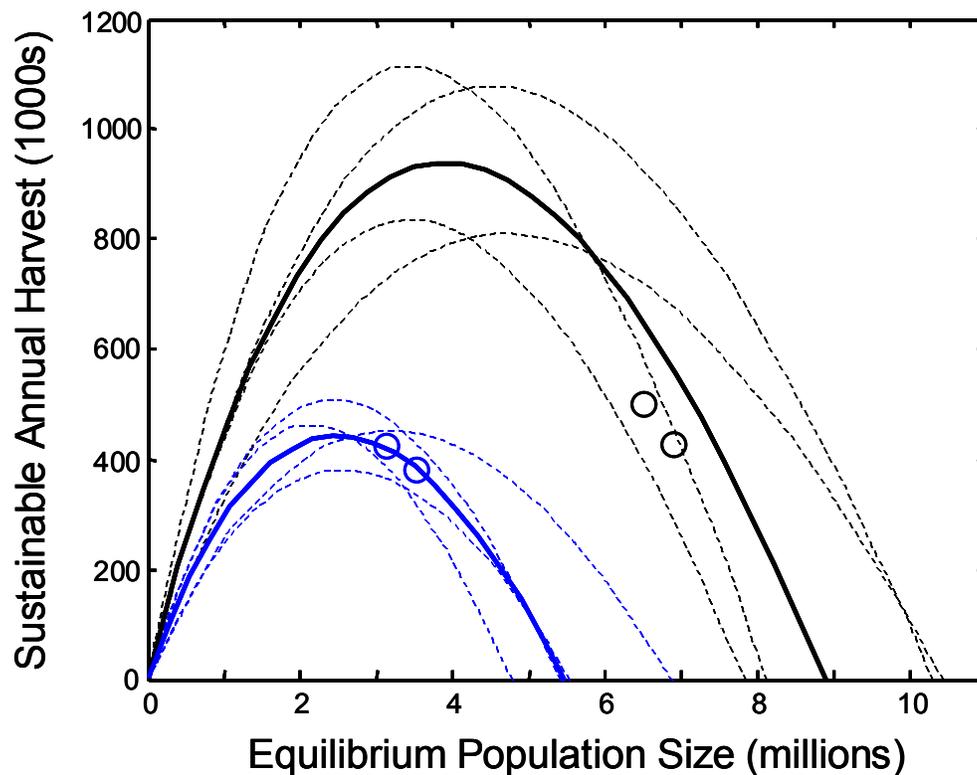


Fig. 8.4. Harvest yield curves, pre-1975 (black) and post-1975 (blue), for 5 alternative models. The x-axis is on the observed BPOP scale. The model formed from the maximum likelihood estimate for the recruitment parameters is shown with a bold line. Four alternative models derived from the 80% confidence ellipsoid for the recruitment parameters are shown with dashed lines. The four circles are described in Section 9.3.

models are taken from the 80% confidence ellipsoid of the parameter estimates for the recruitment relationship; models 3 and 4 are the extremes on that ellipsoid with regard to the latitude effect. The set of models shows that there is considerable uncertainty about the estimates of carrying capacity (both pre- and post-1975), as well as the estimates of optimal equilibrium population size and maximum sustainable harvest. Nevertheless, the shift in average latitude produces a substantial decrease in the productive capacity, hence the carrying capacity and the maximum sustainable harvest (Table 8.2). Across all models considered, the shift in average latitude results in a 31.1-46.2% decrease in carrying capacity, and a 39.1-64.7% decrease in sustainable harvest. Thus, even considering the uncertainty in the recruitment relationship, there is strong evidence that the carrying capacity and the sustainable yield have decreased.

Table 8.2. Equilibrium harvest properties for 5 pintail models, contrasting pre-1975 and post-1975 conditions. Quantities shown include: estimated carrying capacity (K), optimal equilibrium population size (N_{eq}^*), maximum sustainable harvest (H_{eq}^*), the change in carrying capacity pre-1975 to post-1975 (ΔK), and the change in sustainable harvest (ΔH).

Model	K		N_{eq}^*		H_{eq}^*		ΔK	ΔH
	Pre75	Post75	Pre75	Post75	Pre75	Post75		
0 (MLE)	8.90	5.47	3.89	2.50	0.941	0.444	-38.6%	-52.8%
1	7.84	5.41	3.43	2.43	0.835	0.509	-31.1	-39.1
2	10.30	5.54	4.51	2.55	1.080	0.382	-46.2	-64.7
3	10.43	6.87	4.69	3.19	0.811	0.453	-34.1	-44.1
4	8.13	4.77	3.43	2.13	1.115	0.463	-41.4	-58.5

There are other sources of uncertainty besides the parametric uncertainty in the recruitment relationship. First, the functional form of the recruitment relationship can affect harvest dynamics (Runge and Johnson 2002); we have not yet explored different recruitment functions with this pintail model, but we believe the parametric uncertainty analysis above captures the likely range of results that alternative forms would produce. Second, the functional form of the survival relationship can also affect harvest dynamics. This population model, like the model from the original interim pintail harvest strategy, assumes an additive effect of harvest mortality. We have not explored alternative models for harvest mortality. We believe, however, that reasonable alternatives would likely serve to accentuate the differences pre- and post-1975. If harvest mortality is compensatory with natural mortality, the most plausible mechanism is through density-dependent winter survival. Thus, the compensatory effect should increase as a population approaches its winter carrying capacity and becomes more severely winter limited. In the case here, if changes on the breeding grounds have reduced the (summer) carrying capacity and hence the population size, then, all other things being equal, the population should be less winter limited, and hence, should show more additive effects of harvest mortality. This is because in a less winter-limited setting, there are more resources per individual, competition is reduced, and thus each death stands on its own—if an animal dies, that death does not free up any resources that would improve the survival chances of another animal. This argument lends support to the post-1975 yield curves shown in Figs. 8.3 and 8.4. If the pre-1975 populations were more winter limited, harvest may have been more compensatory, and the yield curves shown may be underestimated. Further exploration of these potential dynamics is warranted, but

at this point, we believe the shift in dynamics as a result of a change in average latitude should be taken at face value.

Note that temporal variation does not produce uncertainty with regard to these curves—the curves are meant to be interpreted as the annual harvest under equilibrium conditions. While this is an abstraction (since a stochastic system is never at equilibrium), it is a fundamentally useful one. We believe we cannot understand the role of annual variation until we understand the deterministic backdrop against which it occurs.

9. Pintail Harvest Strategy

The pintail harvest strategy was modified in July 2004 as follows: “Season closed when the breeding population estimate (BPOP) is less than 1.5 million and the projected Fall Flight is less than 2.0 million. Partial season (restrictive alternative) when the BPOP or Fall Flight exceeds the closure level but the BPOP is less than 2.5 million and projections in the strategy predict a decline in the following year’s BPOP (not including a 6% growth factor). Full season, minimum 1-bird daily bag limit when the BPOP exceeds 2.5 million, regardless of the following year’s BPOP projection. All other existing provisions of the strategy continue to apply” (69 FR 52131). The purpose of this section is to investigate the properties of that strategy. We considered the technical modifications described above, but otherwise took the policy as it was. At this point, we have not looked at major changes, either in the general model structure or in the policy itself.

9.1. State-dependent Harvest Policy

Incorporating the three technical improvements described above into the population model, we can calculate and depict this harvest strategy as in Fig. 9.1. The season would be closed when the *observed* BPOP is less than ~1 million (which is roughly equivalent to a corrected fall flight

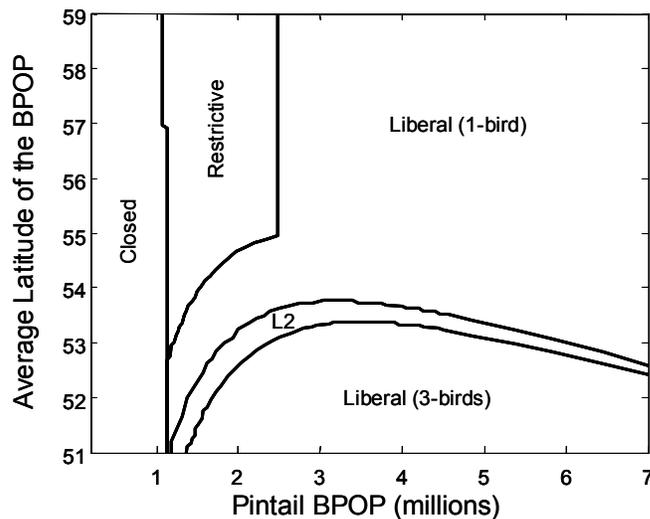


Fig. 9.1. Pintail harvest strategy, based on the July 2004 modifications and three technical improvements (overflight bias correction, new recruitment model, updated harvest models). For a given value of the *observed* (not corrected) pintail BPOP and average latitude of the BPOP, the resulting regulatory decision is shown. Note that this graph assumes the AHM package is “liberal”, thus the “restrictive” regulation implies a season-within-a-season.

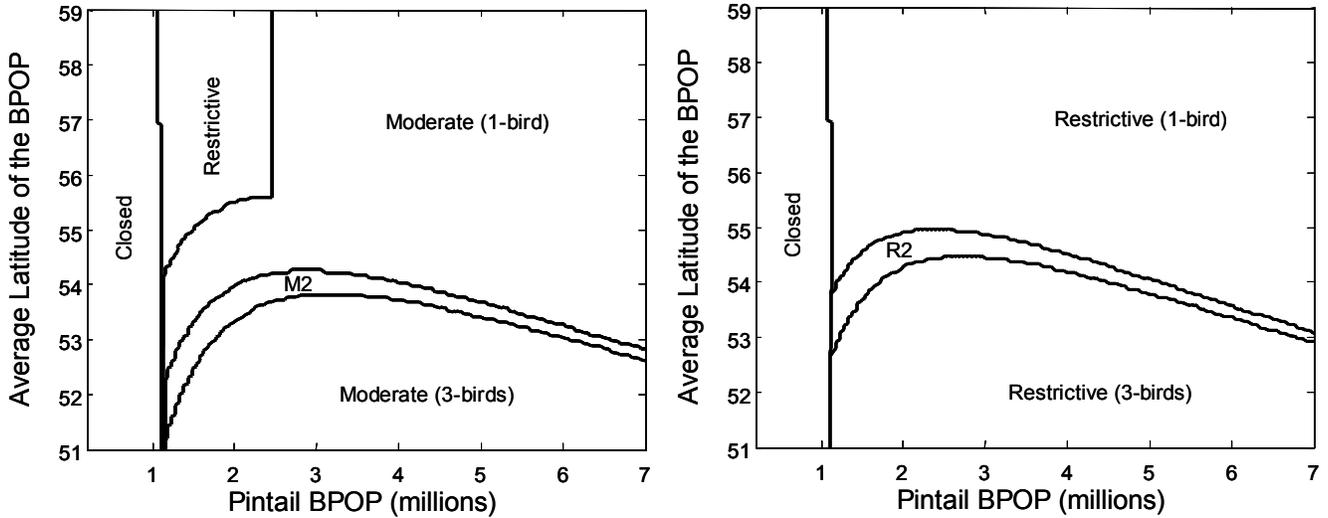


Fig. 9.2. State-dependent pintail harvest strategy, given (A) a moderate AHM package and (B) a restrictive AHM package. For details, see Fig. 9.1.

of 2 million), restrictive when the observed BPOP is less than 2.5 million with a high latitude of the BPOP, and liberal otherwise (this graph assumes the AHM package is liberal; the restrictive section of the graph implies a season-within-a-season). More than one bird in the bag is allowed when the population growth is expected to be greater than 6%. The corresponding state-dependent harvest policies when the AHM package is moderate or restrictive are shown in Fig. 9.2. When the AHM package is moderate, a restrictive season-within-a-season is possible. When the AHM package is restrictive, the pintail season is either also restrictive, or else closed. Bag limits greater than 1 are possible if the population growth is expected to be greater than 6%.

9.2. Comparison to Past Policy Decisions

The past eight years of pintail regulations are shown in Table 9.1, along with the regulations that would be called for under the harvest strategy as modified in 2004, with and without the

Table 9.1. Regulations under the interim pintail harvest strategy, compared to regulations that would be have been called for using several modifications. “Regulation” gives the actual regulations enacted; “2004 Strategy” indicates what would be called for under the 2004 modifications to the pintail harvest strategy using the *old* population model; “2004+” indicates what would be called for under the 2004 strategy using the *new* population model (i.e., the technical modifications described above). “L3” refers to a liberal season length with three birds in the bag for all four flyways. “L3112” refers to a liberal season length with different bag limits in the four flyways (reading from west to east).

Year	BPOP (obs)	Latitude	Regulation	2004 Strategy	2004+
1997	3.56	52.7	L3	L3112	L3
1998	2.52	55.4	L1	L1	L1
1999	3.06	54.6	L1	L1	L1
2000	2.91	57.7	L1	L1	L1
2001	3.30	55.8	L1	L1	L1
2002	1.79	57.9	R1	R1	R1
2003	2.56	55.3	R1	L1	L1
2004	2.18	56.5	R1	R1	R1

additional technical modifications described above. There are only two noteworthy differences in the table. First, the 2003 pintail season length would have been liberal, not season-within-a-season, under the 2004 strategy, because the observed population size was just above the threshold (2.5 million) that precludes a season-within-a-season. Second, the 1997 bag limits would have differed among flyways under the 2004 strategy but not the 2004+ strategy, because the BPOP correction in the 2004+ strategy predicts a larger allowable harvest. The harvest models were modified in 2002, which is why the actual regulation in 1997 differs from what's called for under the 2004 strategy.

9.3. Predicted Performance of the Harvest Strategy

To investigate the predicted future performance of the pintail harvest strategy, we had to simulate the mid-continent mallard and pintail dynamics together, because the same variable—water on the prairies, either as measured by Canadian Ponds or pintail latitude—drives the dynamics of both species, and we wanted to be sure we had the right correlation structure for pintail latitude and AHM regulations. To do this, we used the existing AHM models, with the 2004 model weights (USFWS 2004) to simulate the number of Canadian Ponds and the corresponding AHM package under two scenarios: with and without inclusion of the North American Waterfowl Management Plan goal for mid-continent mallards in the AHM objective function. Next, we used equations (7.1) and (7.2) to generate corresponding values for Latitude of the pintail population under pre-1975 and post-1975 environmental conditions. Finally, we used the pintail population model described herein, with the updated harvest strategy (the “2004+” strategy), to simulate pintail dynamics and harvest.

With the North American goal included in the objective function (i.e., current policy) and with pre-1975 environmental conditions, pintail season lengths followed AHM regulations almost exactly (Table 9.2). If the AHM season length was liberal, the pintail season length was liberal;

Table 9.2. Predictive performance of the pintail harvest strategy under four scenarios: under pre-1975 and post-1975 environmental conditions; and with and without the North American Waterfowl Management Plan goal in the mid-continent mallard AHM objective function. The frequencies of the AHM packages and pintail regulations are shown; for the pintail regulations, only the season length is shown, different bag limits are lumped together. The observed pintail breeding population size and annual harvest are expressed in millions.

	With MCM NA goal			Without MCM NA goal		
	AHM Pkg	NOPI Pre	NOPI Post	AHM Pkg	NOPI Pre	NOPI Post
<i>Regulation Frequency</i>						
Closed	12 %	12 %	19 %	14 %	14 %	22 %
Restrictive	39 %	39 %	36 %	6 %	6 %	5 %
R-SIS	--	0.1 %	6 %	--	0.2 %	18 %
Moderate	8 %	8 %	6 %	6 %	6 %	4 %
Liberal	41 %	41 %	33 %	73 %	73 %	50 %
<i>Average Population Size and Yield</i>						
Mean BPOP (sd)	--	6.9 (1.7)	3.7 (1.9)	--	6.6 (1.6)	3.2 (1.7)
Mean Harv. (sd)	--	0.43 (0.19)	0.38 (0.22)	--	0.50 (0.21)	0.43 (0.24)

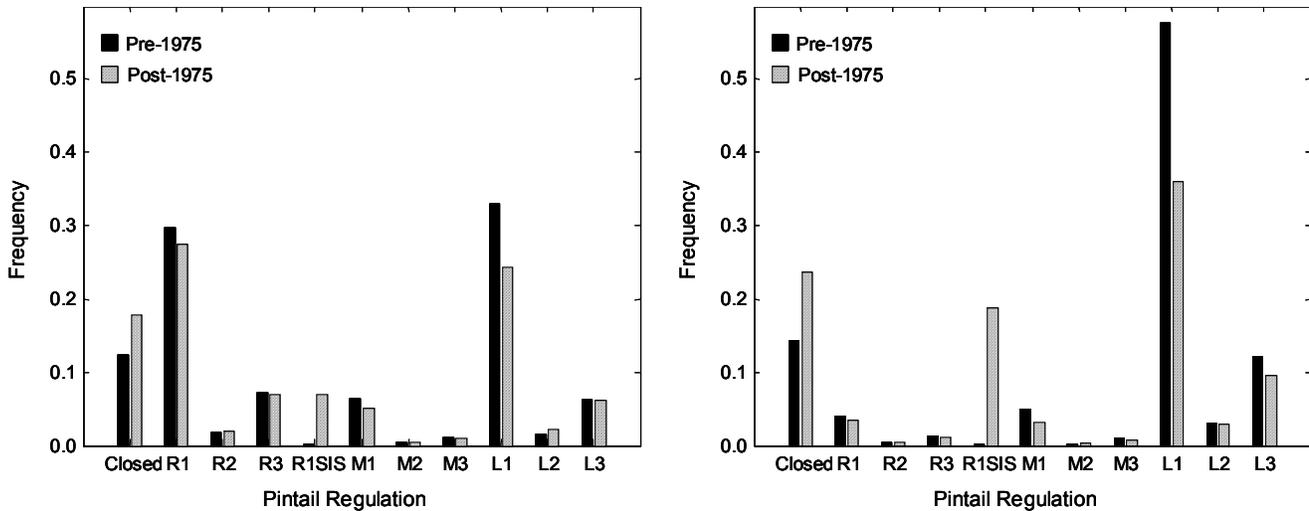


Fig. 9.3. Expected frequencies of pintail regulations under the 2004 harvest strategy, including the proposed technical modifications. Solid bars indicate pre-1975 environmental conditions (average pintail latitude); hatched bars indicate post-1975 conditions. The regulations are indicated with a season length and bag limit (i.e., “R3” indicates a restrictive pintail season length with 3 birds in the bag). “R1SIS” denotes a restrictive pintail season within a longer duck season. (a) Left graph shows scenarios in which the North American goal is included in the AHM objective function. (b) Right graph shows scenarios in which the NA goal is not included.

when the AHM season was moderate, the pintail season was moderate, etc.; although there could be variation in pintail bag limit (Fig. 9.3a). Under this scenario, the expected long-term average pintail breeding population size was 6.9 million (on the observed scale), and the average annual harvest was 430,000 (this point is plotted on Fig. 8.4 for reference to the equilibrium yield curve). Since this point is far below the maximum sustained yield, this pintail policy is considerably more restrictive than what could have been supported, and indeed, what was practiced, prior to the mid-1970s. This conservatism is a result of the requirement for 6% growth to allow bag limits larger than 1.

Under post-1975 environmental conditions (but still including the North American goal), the pintail regulations begin to deviate somewhat from the AHM packages, with an increased frequency of restrictive (season-within-a-season) and closed pintail regulations, and a decreased frequency of liberal pintail regulations. Still, the expected deviation is not large—overall, the season-within-a-season (R-SIS) was only selected 6% of the time, and a closed pintail season with an open duck season occurred only 7% of the time (Table 9.2, Fig. 9.3a). The average pintail population size showed the effects of the loss of productivity, decreasing to 3.7 million (on the observed scale). The average annual harvest, at 380,000, however, did not decrease by much. Plotting this point on the equilibrium yield curve (Fig. 8.4) indicates that this policy is expected to manage the population closer to, but less than, the maximum sustained yield (444,000, see Table 8.2).

When the North American goal is not included in the AHM objective function, both the mid-continent mallard and pintail regulations become more liberal, and the effect of the change in environmental conditions is stronger. Again, under pre-1975 conditions, the pintail regulations track the AHM regulations very closely (Table 9.2, Fig. 9.3b). The expected pintail population size (6.6 million) is slightly lower than when the NA goal is included, and the harvest is higher (500,000, Table 9.2). Under post-1975 conditions, the restrictive season-within-a-season is

called for 18% of the time, and closed pintail season during and open duck season occur 8% of the time, while the frequency of liberal pintail seasons drops to 50%. Without the NA goal in the AHM objective, the pintail policy manages the population very close to the maximum sustained yield (see point on Fig. 8.4).

10. Conclusion

10.1. Summary

We believe that the technical improvements suggested above could be used to improve the predictive models used to assess northern pintail dynamics and inform harvest regulations. Most importantly, the overflight-bias correction (a) provides a way to remove the jumps in observed pintail population sizes brought about by changes in distribution rather than abundance; and (b) substantially improves the predictive ability of the population model. The other proposed modifications to the recruitment and harvest models are fairly routine updates. The end result is that technical improvements in model form, plus longer data sets, add up to a better predictive model for pintail dynamics.

This modeling structure allowed us to understand the implications of an apparent change in pintail dynamics in the mid-1970s. The latitude variable provides an important link that ties together pintail settling patterns and behavior with the demographic consequences. The shift to a more northward distribution, even after controlling for water availability, explains a decrease in reproduction at the continental level and predicts loss of carrying capacity and harvest potential. While the evidence suggests that the drop in pintail numbers has been driven by these changes in recruitment, not by harvest, there is nevertheless a real consequence for future harvest potential.

The existing harvest strategy, evaluated with the proposed technical modifications, (1) appears to be responsive to the loss of harvest potential by providing levels of harvest that are sustainable under the new system dynamics; (2) appears to be somewhat conservative in not trying to harvest at the maximum sustainable level; (3) is predicted to provide restrictive seasons-within-a-season fairly seldom, instead mostly following the AHM season length; and (4) is predicted to mostly call for 1-bird bag limits for pintails. The incorporation of the North American goal in the AHM objective function appears to be conferring some protection to pintails, in holding the pintail population size somewhat higher, in reducing the level of harvest, and in decreasing the frequency of seasons-within-a-season.

10.2. Future Steps

There are a series of additional analyses that we can envision. First, we are interested in whether a balance-equation approach, developed like the mid-continent mallard models (Runge et al. 2002), would shed additional light on pintail dynamics. At least, a broader set of considerations concerning the model structure, including such things as a compensatory harvest mortality hypothesis, could be entertained. Second, it is possible to derive an optimal harvest strategy through methods like stochastic dynamic programming, provided that an explicit harvest management objective could be specified. Comparison of such an optimal policy to the current harvest strategy might help us understand the potential performance of the current strategy.

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13. Appendices

Appendix 1. May aerial survey estimates for pintails. The corrected BPOP is calculated using equation 5.4.

Year	BPOP	SE of BPOP	Latitude	Longitude	BPOP (corrected) millions
1960	5,722,160	323,233	53.78	114.60	6.6821
1961	4,218,159	496,169	55.77	118.22	5.7106
1962	3,623,524	243,150	54.23	118.44	4.3760
1963	3,846,015	255,565	54.18	118.91	4.6286
1964	3,291,227	239,393	54.73	119.95	4.1263
1965	3,591,918	221,851	53.78	116.47	4.1966
1966	4,811,934	265,557	52.62	114.94	5.1563
1967	5,277,693	341,940	52.14	113.73	5.4605
1968	3,489,395	244,623	56.61	128.25	5.0264
1969	5,903,888	296,191	51.68	111.91	5.9039
1970	6,391,987	396,727	51.96	112.36	6.5263
1971	5,847,204	368,133	51.74	111.07	5.8711
1972	6,978,954	364,513	52.69	113.89	7.5175
1973	4,356,220	266,988	54.76	121.42	5.4725
1974	6,598,182	345,810	53.06	115.32	7.3078
1975	5,900,370	267,326	52.58	113.17	6.3076
1976	5,475,644	299,161	53.23	115.29	6.1410
1977	3,926,093	246,802	58.86	133.74	6.6807
1978	5,108,179	267,756	52.69	116.60	5.5034
1979	5,376,133	274,413	52.33	114.31	5.6388
1980	4,508,077	228,608	56.94	127.53	6.6533
1981	3,479,479	260,506	57.59	130.29	5.3902
1982	3,708,758	226,559	53.06	116.99	4.1071
1983	3,510,642	178,068	55.05	120.27	4.5067
1984	2,964,801	166,782	56.03	123.82	4.0911
1985	2,515,493	142,969	55.14	121.54	3.2496
1986	2,739,747	152,100	54.85	122.08	3.4649
1987	2,628,344	159,412	55.83	126.06	3.5743
1988	2,005,522	164,048	59.06	136.70	3.4654
1989	2,111,902	181,253	56.88	128.87	3.1039
1990	2,256,630	183,280	57.81	132.69	3.5525
1991	1,803,385	131,280	59.81	136.31	3.2920
1992	2,098,139	160,972	57.66	130.02	3.2668
1993	2,053,418	124,184	55.95	126.46	2.8161
1994	2,972,266	188,005	53.56	120.31	3.4165
1995	2,757,866	177,594	53.73	119.92	3.2093
1996	2,735,863	147,542	53.51	120.61	3.1306
1997	3,557,991	194,237	52.71	116.09	3.8385
1998	2,520,649	136,806	55.39	125.77	3.3180
1999	3,057,888	230,532	54.61	123.58	3.7985
2000	2,907,559	170,467	57.74	131.56	4.5542
2001	3,295,994	266,584	55.85	127.25	4.4895
2002	1,789,710	125,199	57.91	134.13	2.8389
2003	2,558,229	174,797	55.29	124.72	3.3409
2004	2,184,602	155,233	56.50	129.00	3.1211

Appendix 2. Direct recovery rates of pintails.

Year	Young Male		Adult Male		Young Female		Adult Female	
	f	SE	f	SE	f	SE	f	SE
1960	0.0590	0.0039	0.0250	0.0021	0.0455	0.0035	0.0217	0.0032
1961	0.0449	0.0044	0.0223	0.0018	0.0302	0.0037	0.0135	0.0019
1962	0.0376	0.0047	0.0209	0.0025	0.0291	0.0042	0.0227	0.0037
1963	0.0436	0.0044	0.0202	0.0026	0.0339	0.0038	0.0185	0.0037
1964	0.0385	0.0033	0.0246	0.0024	0.0373	0.0033	0.0281	0.0031
1965	0.0369	0.0032	0.0271	0.0033	0.0264	0.0026	0.0181	0.0030
1966	0.0449	0.0026	0.0206	0.0016	0.0329	0.0021	0.0162	0.0018
1967	0.0396	0.0023	0.0236	0.0018	0.0307	0.0018	0.0215	0.0020
1968	0.0382	0.0041	0.0177	0.0014	0.0236	0.0033	0.0140	0.0019
1969	0.0408	0.0028	0.0285	0.0035	0.0338	0.0026	0.0231	0.0030
1970	0.0432	0.0033	0.0275	0.0026	0.0415	0.0032	0.0256	0.0029
1971	0.0433	0.0035	0.0267	0.0025	0.0293	0.0029	0.0180	0.0023
1972	0.0379	0.0035	0.0292	0.0024	0.0308	0.0034	0.0162	0.0024
1973	0.0291	0.0039	0.0150	0.0019	0.0251	0.0036	0.0131	0.0022
1974	0.0334	0.0025	0.0156	0.0022	0.0276	0.0023	0.0159	0.0026
1975	0.0342	0.0029	0.0291	0.0028	0.0245	0.0025	0.0156	0.0024
1976	0.0316	0.0028	0.0224	0.0023	0.0295	0.0027	0.0162	0.0023
1977	0.0289	0.0034	0.0171	0.0016	0.0237	0.0031	0.0131	0.0017
1978	0.0342	0.0029	0.0214	0.0023	0.0228	0.0025	0.0137	0.0025
1979	0.0405	0.0026	0.0191	0.0022	0.0239	0.0021	0.0137	0.0022
1980	0.0397	0.0040	0.0186	0.0024	0.0270	0.0034	0.0125	0.0023
1981	0.0334	0.0048	0.0180	0.0034	0.0245	0.0041	0.0116	0.0028
1982	0.0363	0.0049	0.0212	0.0038	0.0154	0.0031	0.0159	0.0030
1983	0.0318	0.0034	0.0187	0.0031	0.0295	0.0030	0.0134	0.0026
1984	0.0239	0.0038	0.0135	0.0026	0.0244	0.0037	0.0135	0.0027
1985	0.0262	0.0038	0.0112	0.0028	0.0199	0.0032	0.0065	0.0020
1986	0.0255	0.0031	0.0187	0.0029	0.0175	0.0025	0.0090	0.0021
1987	0.0282	0.0042	0.0161	0.0024	0.0161	0.0032	0.0084	0.0020
1988	0.0175	0.0050	0.0075	0.0014	0.0092	0.0035	0.0057	0.0014
1989	0.0203	0.0037	0.0052	0.0016	0.0114	0.0028	0.0082	0.0021
1990	0.0195	0.0032	0.0069	0.0013	0.0111	0.0023	0.0067	0.0015
1991	0.0179	0.0031	0.0073	0.0017	0.0126	0.0027	0.0047	0.0013
1992	0.0272	0.0037	0.0124	0.0015	0.0215	0.0031	0.0062	0.0012
1993	0.0262	0.0033	0.0120	0.0021	0.0163	0.0025	0.0071	0.0018
1994	0.0194	0.0023	0.0124	0.0021	0.0159	0.0020	0.0078	0.0018
1995	0.0284	0.0033	0.0188	0.0022	0.0239	0.0028	0.0129	0.0020
1996	0.0335	0.0036	0.0124	0.0018	0.0197	0.0026	0.0090	0.0015
1997	0.0460	0.0033	0.0250	0.0032	0.0328	0.0024	0.0145	0.0020
1998	0.0535	0.0046	0.0305	0.0025	0.0352	0.0036	0.0129	0.0017
1999	0.0618	0.0056	0.0334	0.0034	0.0500	0.0044	0.0232	0.0026
2000	0.0666	0.0061	0.0313	0.0025	0.0476	0.0048	0.0203	0.0020
2001	0.0643	0.0063	0.0278	0.0026	0.0466	0.0047	0.0154	0.0019
2002	0.0436	0.0052	0.0191	0.0029	0.0416	0.0046	0.0137	0.0024
2003	0.0426	0.0036	0.0174	0.0029	0.0303	0.0029	0.0116	0.0020

Appendix 3. Pintail relative vulnerability to the US harvest, and female harvest and population age ratios.

Year	Male Vulnerability	SE	Female Vulnerability	SE	Female harvest age ratio	Female population age ratio
1961	2.0137	0.2550	2.2286	0.4184	1.0981877	0.4927607
1962	1.8026	0.3100	1.2821	0.2803	1.4467156	1.1284102
1963	2.1545	0.3508	1.8316	0.4176	2.0222369	1.1040903
1964	1.5653	0.2036	1.3247	0.1862	1.62864	1.229418
1965	1.3640	0.2054	1.4540	0.2783	3.1222932	2.1474382
1966	2.1780	0.2125	2.0300	0.2619	2.1198433	1.0442514
1967	1.6743	0.1615	1.4278	0.1599	2.4251812	1.6985448
1968	2.1558	0.2879	1.6803	0.3266	1.0962068	0.6523926
1969	1.4298	0.2033	1.4648	0.2214	2.9874863	2.0394568
1970	1.5671	0.1914	1.6180	0.2219	1.9900335	1.2299615
1971	1.6236	0.2010	1.6266	0.2627	1.5003959	0.9223856
1972	1.2999	0.1633	1.9003	0.3519	1.5926621	0.8381097
1973	1.9376	0.3527	1.9183	0.4225	1.2150599	0.6334104
1974	2.1427	0.3445	1.7413	0.3217	2.1053845	1.2090873
1975	1.1771	0.1513	1.5699	0.2859	1.7213364	1.0964296
1976	1.4138	0.1887	1.8174	0.3028	1.693397	0.9317625
1977	1.6911	0.2506	1.8131	0.3381	1.0355985	0.5711825
1978	1.5949	0.2200	1.6630	0.3544	2.4134554	1.4512408
1979	2.1239	0.2785	1.7486	0.3161	2.0022622	1.1450421
1980	2.1339	0.3468	2.1531	0.4735	1.0340407	0.4802499
1981	1.8571	0.4449	2.1179	0.6211	1.2394886	0.5852458
1982	1.7099	0.3854	0.9729	0.2685	1.5456639	1.5887206
1983	1.7009	0.3389	2.2035	0.4788	1.8597628	0.8439904
1984	1.7664	0.4433	1.8069	0.4505	1.620685	0.8969284
1985	2.3487	0.6747	3.0769	1.0885	1.9554672	0.6355218
1986	1.3688	0.2681	1.9381	0.5208	2.1429956	1.1057032
1987	1.7485	0.3666	1.9189	0.5997	1.4420961	0.7515389
1988	2.3213	0.7870	1.6195	0.7243	1.1845318	0.7314347
1989	3.8787	1.3670	1.3892	0.4897	2.1369003	1.5381712
1990	2.8063	0.6893	1.6545	0.4972	2.1245967	1.2841516
1991	2.4506	0.7058	2.7124	0.9452	1.6749028	0.6175086
1992	2.1940	0.4019	3.4577	0.8510	2.0077461	0.5806563
1993	2.1909	0.4708	2.3053	0.6718	2.3789047	1.0319181
1994	1.5614	0.3296	2.0435	0.5435	2.7228894	1.3324599
1995	1.5128	0.2492	1.8503	0.3646	1.7939586	0.9695714
1996	2.7069	0.4881	2.1719	0.4642	2.0975601	0.9657527
1997	1.8442	0.2714	2.2642	0.3569	3.0646875	1.353515
1998	1.7555	0.2071	2.7182	0.4475	1.7321089	0.6372258
1999	1.8504	0.2503	2.1571	0.3107	2.0450855	0.9480774
2000	2.1249	0.2559	2.3457	0.3320	1.4341233	0.6113802
2001	2.3171	0.3149	3.0202	0.4760	1.8863663	0.6245928
2002	2.2820	0.4414	3.0396	0.6254	2.0190861	0.6642518
2003	2.4503	0.4547	2.6092	0.5095	2.9889352	1.1455408

Appendix 4. Adult male pintail annual survival and recovery rates.

Male Year	Adult		95% CI		Adult f	95% CI		
	S	SE	Lower	Upper		SE	Lower	Upper
1960	0.7013	0.0454	0.6055	0.7822	0.0266	0.0022	0.0227	0.0312
1961	0.6669	0.0448	0.5743	0.7482	0.0234	0.0015	0.0206	0.0264
1962	0.8892	0.0713	0.6602	0.9707	0.0208	0.0016	0.0180	0.0241
1963	0.6506	0.0511	0.5452	0.7431	0.0264	0.0019	0.0229	0.0305
1964	0.6749	0.0540	0.5617	0.7708	0.0262	0.0017	0.0231	0.0297
1965	0.8507	0.0623	0.6854	0.9371	0.0237	0.0018	0.0203	0.0275
1966	0.7704	0.0428	0.6762	0.8435	0.0236	0.0012	0.0213	0.0262
1967	0.6996	0.0383	0.6197	0.7689	0.0263	0.0013	0.0238	0.0290
1968	0.5976	0.0430	0.5112	0.6784	0.0216	0.0010	0.0196	0.0237
1969	0.7285	0.0566	0.6049	0.8247	0.0265	0.0019	0.0231	0.0305
1970	0.7869	0.0531	0.6649	0.8729	0.0347	0.0019	0.0312	0.0386
1971	0.6472	0.0425	0.5601	0.7254	0.0280	0.0016	0.0250	0.0313
1972	0.8216	0.0545	0.6897	0.9051	0.0295	0.0016	0.0265	0.0328
1973	0.6944	0.0516	0.5852	0.7854	0.0223	0.0014	0.0198	0.0251
1974	0.6847	0.0518	0.5757	0.7766	0.0272	0.0018	0.0239	0.0309
1975	0.7472	0.0530	0.6303	0.8367	0.0311	0.0019	0.0277	0.0350
1976	0.7409	0.0473	0.6383	0.8225	0.0287	0.0017	0.0256	0.0321
1977	0.7151	0.0472	0.6146	0.7981	0.0227	0.0012	0.0205	0.0252
1978	0.7459	0.0555	0.6232	0.8390	0.0248	0.0015	0.0220	0.0280
1979	0.8651	0.0724	0.6552	0.9559	0.0253	0.0016	0.0224	0.0286
1980	0.6472	0.0733	0.4943	0.7748	0.0226	0.0016	0.0196	0.0259
1981	0.6184	0.0783	0.4581	0.7564	0.0180	0.0018	0.0147	0.0220
1982	0.9837	0.1228	0.0000	1.0000	0.0258	0.0025	0.0214	0.0311
1983	0.6457	0.0812	0.4761	0.7852	0.0216	0.0021	0.0178	0.0261
1984	0.7608	0.1040	0.5093	0.9070	0.0188	0.0019	0.0154	0.0228
1985	0.6500	0.0862	0.4691	0.7961	0.0180	0.0021	0.0143	0.0226
1986	0.7391	0.0807	0.5550	0.8656	0.0179	0.0017	0.0148	0.0216
1987	0.8277	0.0835	0.6040	0.9380	0.0209	0.0018	0.0176	0.0247
1988	0.6754	0.0730	0.5201	0.7998	0.0095	0.0009	0.0079	0.0115
1989	0.8726	0.0906	0.5809	0.9713	0.0117	0.0013	0.0094	0.0144
1990	0.6807	0.0640	0.5447	0.7916	0.0120	0.0010	0.0102	0.0142
1991	0.7556	0.0649	0.6082	0.8603	0.0108	0.0011	0.0089	0.0131
1992	0.7726	0.0612	0.6318	0.8705	0.0144	0.0010	0.0125	0.0166
1993	0.8334	0.0757	0.6323	0.9357	0.0155	0.0013	0.0132	0.0182
1994	0.8047	0.0681	0.6379	0.9060	0.0161	0.0013	0.0137	0.0189
1995	0.8265	0.0650	0.6621	0.9205	0.0195	0.0013	0.0171	0.0223
1996	0.5919	0.0499	0.4918	0.6850	0.0170	0.0012	0.0148	0.0196
1997	0.7564	0.0580	0.6263	0.8519	0.0260	0.0019	0.0225	0.0300
1998	0.7436	0.0561	0.6196	0.8378	0.0325	0.0017	0.0292	0.0361
1999	0.8206	0.0670	0.6522	0.9177	0.0337	0.0023	0.0295	0.0384
2000	0.7035	0.0619	0.5700	0.8094	0.0296	0.0018	0.0264	0.0333
2001	0.8539	0.1232	0.4577	0.9759	0.0289	0.0021	0.0250	0.0333
2002	0.6227	0.1288	0.3604	0.8286	0.0174	0.0022	0.0135	0.0223
2003					0.0189	0.0030	0.0138	0.0257

Appendix 5. Young male pintail annual survival and recovery rates.

Male Year	Young		95% CI		Young f	95% CI		
	S	SE	Lower	Upper		SE	Lower	Upper
1960	0.4459	0.0403	0.3689	0.5255	0.0670	0.0042	0.0593	0.0756
1961	0.4190	0.0482	0.3285	0.5152	0.0517	0.0047	0.0432	0.0618
1962	0.6285	0.0747	0.4746	0.7601	0.0426	0.0050	0.0338	0.0536
1963	0.5202	0.0555	0.4122	0.6263	0.0519	0.0048	0.0433	0.0622
1964	0.5649	0.0530	0.4597	0.6646	0.0493	0.0037	0.0424	0.0571
1965	0.6105	0.0484	0.5126	0.7002	0.0533	0.0038	0.0463	0.0612
1966	0.6383	0.0416	0.5536	0.7152	0.0582	0.0029	0.0528	0.0642
1967	0.6384	0.0382	0.5608	0.7094	0.0516	0.0026	0.0468	0.0569
1968	0.5192	0.0560	0.4103	0.6264	0.0509	0.0047	0.0425	0.0609
1969	0.5633	0.0406	0.4827	0.6407	0.0509	0.0031	0.0451	0.0575
1970	0.6274	0.0507	0.5240	0.7203	0.0626	0.0040	0.0553	0.0708
1971	0.6046	0.0484	0.5070	0.6945	0.0571	0.0040	0.0497	0.0655
1972	0.6323	0.0571	0.5151	0.7357	0.0530	0.0042	0.0455	0.0618
1973	0.6135	0.0662	0.4787	0.7328	0.0458	0.0049	0.0372	0.0564
1974	0.5419	0.0420	0.4593	0.6224	0.0440	0.0029	0.0386	0.0500
1975	0.5799	0.0476	0.4849	0.6693	0.0478	0.0034	0.0416	0.0549
1976	0.6216	0.0480	0.5240	0.7102	0.0470	0.0033	0.0409	0.0540
1977	0.6349	0.0614	0.5084	0.7451	0.0426	0.0040	0.0353	0.0512
1978	0.5772	0.0508	0.4758	0.6725	0.0439	0.0033	0.0378	0.0508
1979	0.7801	0.0658	0.6259	0.8827	0.0553	0.0030	0.0497	0.0615
1980	0.6996	0.0898	0.5019	0.8433	0.0517	0.0046	0.0434	0.0614
1981	0.6001	0.0855	0.4274	0.7510	0.0421	0.0054	0.0327	0.0541
1982	0.7317	0.1083	0.4804	0.8895	0.0458	0.0055	0.0362	0.0578
1983	0.6719	0.0846	0.4911	0.8129	0.0425	0.0039	0.0354	0.0510
1984	0.6753	0.1070	0.4442	0.8441	0.0349	0.0045	0.0270	0.0450
1985	0.6964	0.0933	0.4913	0.8449	0.0324	0.0042	0.0251	0.0417
1986	0.8434	0.0923	0.5780	0.9549	0.0358	0.0037	0.0292	0.0437
1987	0.9537	0.1225	0.0825	0.9998	0.0390	0.0049	0.0305	0.0498
1988	0.7934	0.1465	0.3999	0.9568	0.0262	0.0061	0.0166	0.0413
1989	0.6855	0.0966	0.4753	0.8398	0.0344	0.0048	0.0261	0.0452
1990	0.6857	0.0856	0.5003	0.8261	0.0270	0.0038	0.0206	0.0355
1991	0.6799	0.0743	0.5209	0.8057	0.0290	0.0040	0.0222	0.0379
1992	0.6796	0.0755	0.5181	0.8072	0.0379	0.0043	0.0303	0.0474
1993	0.6863	0.0733	0.5287	0.8101	0.0354	0.0039	0.0286	0.0438
1994	0.8141	0.0674	0.6466	0.9129	0.0287	0.0028	0.0236	0.0348
1995	0.7206	0.0721	0.5610	0.8388	0.0384	0.0038	0.0316	0.0467
1996	0.5388	0.0558	0.4293	0.6447	0.0403	0.0039	0.0333	0.0487
1997	0.6244	0.0470	0.5288	0.7112	0.0542	0.0036	0.0476	0.0616
1998	0.5913	0.0619	0.4669	0.7050	0.0621	0.0049	0.0532	0.0724
1999	0.8516	0.0919	0.5797	0.9598	0.0688	0.0059	0.0581	0.0813
2000	0.7521	0.1026	0.5079	0.8992	0.0719	0.0063	0.0605	0.0853
2001	0.9916	0.1937	0.0000	1.0000	0.0749	0.0067	0.0627	0.0892
2002	0.6470	0.1797	0.2816	0.8955	0.0462	0.0053	0.0368	0.0578
2003					0.0519	0.0040	0.0446	0.0602

Appendix 6. Adult female pintail annual survival and recovery rates.

Female Year	Adult		95% CI		Adult f	SE	95% CI	
	S	SE	Lower	Upper			Lower	Upper
1960	0.7662	0.1077	0.5021	0.9141	0.0242	0.0034	0.0183	0.0319
1961	0.5577	0.0761	0.4078	0.6978	0.0152	0.0017	0.0122	0.0188
1962	0.6193	0.0956	0.4236	0.7827	0.0161	0.0021	0.0124	0.0209
1963	0.6100	0.0816	0.4441	0.7539	0.0210	0.0027	0.0163	0.0270
1964	0.6474	0.0790	0.4824	0.7834	0.0281	0.0025	0.0237	0.0333
1965	0.7376	0.0843	0.5449	0.8684	0.0185	0.0021	0.0148	0.0230
1966	0.5492	0.0481	0.4543	0.6406	0.0193	0.0015	0.0166	0.0225
1967	0.6375	0.0574	0.5195	0.7410	0.0234	0.0016	0.0205	0.0268
1968	0.5817	0.0591	0.4634	0.6912	0.0183	0.0014	0.0156	0.0213
1969	0.6427	0.0664	0.5051	0.7602	0.0222	0.0019	0.0187	0.0264
1970	0.7983	0.0829	0.5906	0.9156	0.0278	0.0022	0.0239	0.0324
1971	0.6321	0.0725	0.4825	0.7600	0.0199	0.0017	0.0168	0.0235
1972	0.6578	0.0825	0.4838	0.7978	0.0202	0.0019	0.0168	0.0243
1973	0.5654	0.0741	0.4187	0.7014	0.0164	0.0017	0.0134	0.0200
1974	0.7324	0.0961	0.5114	0.8774	0.0181	0.0019	0.0147	0.0224
1975	0.5008	0.0622	0.3812	0.6203	0.0193	0.0019	0.0159	0.0234
1976	0.7166	0.0814	0.5355	0.8472	0.0195	0.0018	0.0162	0.0234
1977	0.5918	0.0757	0.4395	0.7283	0.0154	0.0014	0.0129	0.0183
1978	0.6604	0.0918	0.4658	0.8127	0.0170	0.0019	0.0136	0.0212
1979	0.6384	0.0891	0.4532	0.7900	0.0152	0.0016	0.0124	0.0188
1980	0.7797	0.1107	0.5003	0.9260	0.0170	0.0019	0.0136	0.0211
1981	0.4359	0.0649	0.3154	0.5645	0.0099	0.0011	0.0079	0.0124
1982	0.7507	0.1245	0.4499	0.9173	0.0177	0.0022	0.0138	0.0227
1983	0.5052	0.0875	0.3396	0.6696	0.0164	0.0021	0.0127	0.0211
1984	0.8483	0.1740	0.2831	0.9875	0.0165	0.0022	0.0127	0.0214
1985	0.4519	0.0929	0.2833	0.6323	0.0097	0.0017	0.0068	0.0137
1986	0.6808	0.1237	0.4113	0.8669	0.0130	0.0018	0.0099	0.0172
1987	0.8337	0.1606	0.3410	0.9798	0.0135	0.0019	0.0102	0.0179
1988	0.4282	0.0825	0.2790	0.5917	0.0058	0.0010	0.0042	0.0080
1989	0.8946	0.1652	0.2148	0.9962	0.0104	0.0017	0.0076	0.0143
1990	0.7449	0.1384	0.4119	0.9241	0.0076	0.0011	0.0057	0.0102
1991	0.5836	0.1010	0.3829	0.7599	0.0054	0.0009	0.0039	0.0075
1992	0.4897	0.0761	0.3457	0.6354	0.0072	0.0010	0.0055	0.0094
1993	0.6496	0.1006	0.4380	0.8151	0.0083	0.0012	0.0062	0.0111
1994	0.8111	0.1146	0.4977	0.9490	0.0105	0.0014	0.0080	0.0136
1995	0.7073	0.0919	0.5030	0.8523	0.0123	0.0014	0.0099	0.0153
1996	0.5984	0.0744	0.4481	0.7322	0.0113	0.0012	0.0092	0.0138
1997	0.6132	0.0709	0.4688	0.7401	0.0124	0.0013	0.0101	0.0152
1998	0.5187	0.0563	0.4091	0.6264	0.0147	0.0013	0.0123	0.0174
1999	0.7499	0.0838	0.5554	0.8780	0.0232	0.0020	0.0195	0.0275
2000	0.5813	0.0714	0.4386	0.7116	0.0197	0.0016	0.0168	0.0232
2001	0.5610	0.0937	0.3775	0.7292	0.0155	0.0015	0.0128	0.0188
2002	0.5689	0.1247	0.3275	0.7814	0.0153	0.0021	0.0116	0.0200
2003					0.0140	0.0022	0.0103	0.0189

Appendix 7. Young female pintail annual survival and recovery rates.

Female Year	Young		95% CI		Young		95% CI	
	S	SE	Lower	Upper	f	SE	Lower	Upper
1960	0.6273	0.0786	0.4656	0.7648	0.0509	0.0037	0.0441	0.0587
1961	0.3916	0.0705	0.2649	0.5348	0.0379	0.0041	0.0307	0.0468
1962	0.4907	0.0873	0.3269	0.6564	0.0384	0.0048	0.0301	0.0490
1963	0.4597	0.0642	0.3390	0.5853	0.0397	0.0041	0.0323	0.0486
1964	0.5124	0.0694	0.3787	0.6444	0.0472	0.0037	0.0405	0.0549
1965	0.5837	0.0634	0.4567	0.7005	0.0370	0.0030	0.0315	0.0434
1966	0.4579	0.0414	0.3786	0.5394	0.0463	0.0025	0.0416	0.0515
1967	0.5648	0.0496	0.4664	0.6584	0.0412	0.0021	0.0372	0.0455
1968	0.5110	0.0705	0.3755	0.6449	0.0339	0.0039	0.0270	0.0426
1969	0.5003	0.0512	0.4014	0.5993	0.0479	0.0031	0.0421	0.0543
1970	0.4547	0.0582	0.3448	0.5692	0.0560	0.0037	0.0492	0.0637
1971	0.6128	0.0780	0.4538	0.7510	0.0425	0.0035	0.0362	0.0498
1972	0.6525	0.0947	0.4529	0.8098	0.0446	0.0041	0.0372	0.0534
1973	0.4259	0.0764	0.2868	0.5779	0.0407	0.0045	0.0327	0.0505
1974	0.6831	0.0813	0.5080	0.8182	0.0396	0.0027	0.0346	0.0453
1975	0.5040	0.0647	0.3797	0.6278	0.0353	0.0030	0.0299	0.0416
1976	0.6612	0.0790	0.4944	0.7957	0.0433	0.0032	0.0374	0.0500
1977	0.6894	0.1047	0.4598	0.8526	0.0347	0.0037	0.0281	0.0427
1978	0.6035	0.0851	0.4312	0.7534	0.0362	0.0032	0.0304	0.0430
1979	0.6146	0.0809	0.4495	0.7570	0.0351	0.0025	0.0306	0.0403
1980	1.0000	0.0001	0.0008	1.0000	0.0335	0.0037	0.0269	0.0416
1981	0.5859	0.1151	0.3583	0.7819	0.0343	0.0048	0.0260	0.0451
1982	0.8214	0.1540	0.3700	0.9730	0.0283	0.0042	0.0211	0.0378
1983	0.5828	0.0985	0.3871	0.7556	0.0396	0.0035	0.0332	0.0471
1984	0.8565	0.1956	0.2088	0.9926	0.0380	0.0046	0.0301	0.0481
1985	0.5096	0.1078	0.3085	0.7077	0.0288	0.0038	0.0221	0.0373
1986	0.7377	0.1335	0.4211	0.9158	0.0265	0.0030	0.0211	0.0331
1987	0.8936	0.2102	0.0992	0.9984	0.0219	0.0037	0.0157	0.0304
1988	0.6538	0.1883	0.2699	0.9061	0.0145	0.0043	0.0080	0.0259
1989	0.6713	0.1630	0.3244	0.8968	0.0249	0.0040	0.0181	0.0341
1990	0.6482	0.1536	0.3298	0.8734	0.0159	0.0027	0.0113	0.0223
1991	0.7683	0.1553	0.3750	0.9482	0.0207	0.0034	0.0149	0.0285
1992	0.4563	0.0930	0.2871	0.6363	0.0279	0.0035	0.0218	0.0358
1993	0.4784	0.0829	0.3235	0.6376	0.0236	0.0030	0.0184	0.0301
1994	0.5868	0.0835	0.4196	0.7361	0.0212	0.0023	0.0171	0.0262
1995	0.6879	0.1015	0.4659	0.8478	0.0339	0.0034	0.0279	0.0411
1996	0.6219	0.0934	0.4303	0.7818	0.0266	0.0031	0.0212	0.0333
1997	0.6557	0.0717	0.5054	0.7802	0.0382	0.0026	0.0334	0.0437
1998	0.4730	0.0691	0.3427	0.6070	0.0413	0.0039	0.0343	0.0495
1999	0.5934	0.0916	0.4094	0.7544	0.0533	0.0045	0.0451	0.0629
2000	0.6991	0.1267	0.4164	0.8832	0.0538	0.0051	0.0446	0.0647
2001	0.8057	0.1689	0.3334	0.9717	0.0557	0.0052	0.0464	0.0667
2002	0.5795	0.1740	0.2537	0.8481	0.0486	0.0050	0.0397	0.0594
2003					0.0382	0.0032	0.0324	0.0450

Appendix 8. Central flyway total pintail harvest and harvest regulations from 1979-2003.

Year	Harvest	HIP	SIS	Days	Bag	DaysBag	MWS
1979	228947	0	0	60	10	600	1697720
1980	193244	0	0	60	10	600	1542961
1981	151023	0	0	60	10	600	349677
1982	158994	0	0	60	10	600	617876
1983	139077	0	0	60	10	600	353175
1984	165804	0	0	60	10	600	593311
1985	83914	0	0	50	3	150	571800
1986	72071	0	0	51	3	153	335335
1987	122420	0	0	51	3	153	516825
1988	36387	0	0	39	1	39	733405
1989	43594	0	0	39	1	39	278559
1990	43206	0	0	39	1	39	589250
1991	28687	0	0	39	1	39	97190
1992	32095	0	0	39	1	39	905320
1993	42274	0	0	39	1	39	319120
1994	61456	0	0	49	1	49	431924
1995	94840	0	0	60	1	60	721753
1996	96634	0	0	60	1	60	317173
1997	189211	0	0	74	3	222	505419
1998	124994	0	0	74	1	74	612712
1999	138397	0	0	74	1	74	992064
2000	136592	0	0	74	1	74	374407
2001	151917	0	0	74	1	74	327797
1999	133317	1	0	74	1	74	992064
2000	134252	1	0	74	1	74	374407
2001	134612	1	0	74	1	74	327797
2002	60407	1	1	39	1	39	492311
2003	55641	1	1	39	1	39	713071
2004							664003

¹HIP = data collected under the HIP program, SIS = years when seasons within a season (partial season) were adopted in the Central Flyway, and MWS is the Midwinter Survey population estimate from the Central flyway.

Appendix 9. Atlantic flyway total pintail harvest and harvest regulations from 1979-2003.

Year	Harvest	HIP	SIS	Days	Bag	DaysBag	MWS
1979	48462	0	0	50	4	200	72071
1980	38869	0	0	50	4	200	54492
1981	27891	0	0	50	4	200	67962
1982	38632	0	0	50	5	250	68900
1983	18636	0	0	50	5	250	48100
1984	34658	0	0	50	5	250	46200
1985	21685	0	0	40	2	80	34000
1986	19033	0	0	40	2	80	47400
1987	15788	0	0	40	1	40	36900
1988	7447	0	0	30	1	30	36235
1989	14588	0	0	30	1	30	55093
1990	10493	0	0	30	1	30	43995
1991	14201	0	0	30	1	30	68937
1992	12470	0	0	30	1	30	42705
1993	12923	0	0	30	1	30	51374
1994	18340	0	0	40	1	40	61510
1995	33163	0	0	50	1	50	50626
1996	19270	0	0	50	1	50	34606
1997	24010	0	0	60	3	180	43165
1998	33594	0	0	60	1	60	45419
1999	29527	0	0	60	1	60	65259
2000	22384	0	0	60	1	60	37890
2001	19950	0	0	60	1	60	47902
1999	25200	1	0	60	1	60	65259
2000	20752	1	0	60	1	60	37890
2001	19276	1	0	60	1	60	47902
2002	17089	1	1	30	1	30	48139
2003	18134	1	1	30	1	30	36324
2004							55523

¹ HIP = data collected under the HIP program, SIS = years when seasons within a season (partial season) were adopted in the Atlantic Flyway, and MWS is the Midwinter Survey population estimate from the Atlantic flyway.

Appendix 10. Mississippi flyway total pintail harvest and harvest regulations from 1979-2003.

Year	Harvest	HIP	SIS	Days	Bag	DaysBag	MWS
1979	213601	0	0	50	10	500	792000
1980	215811	0	0	50	10	500	324025
1981	207864	0	0	50	10	500	699575
1982	126568	0	0	50	10	500	886500
1983	187365	0	0	50	10	500	944975
1984	153680	0	0	50	10	500	532475
1985	124920	0	0	40	3	120	694300
1986	90350	0	0	40	3	120	422675
1987	88305	0	0	40	3	120	250075
1988	39225	0	0	30	1	30	452675
1989	65055	0	0	30	1	30	529275
1990	49487	0	0	30	1	30	650524
1991	40319	0	0	30	1	30	799597
1992	56520	0	0	30	1	30	518954
1993	52635	0	0	30	1	30	202043
1994	81147	0	0	40	1	40	537674
1995	136099	0	0	50	1	50	522198
1996	123817	0	0	50	1	50	631666
1997	144758	0	0	60	3	180	259594
1998	176990	0	0	60	1	60	378870
1999	167666	0	0	60	1	60	316760
2000	161476	0	0	60	1	60	486406
2001	130894	0	0	60	1	60	1207671
1999	148299	1	0	60	1	60	316760
2000	155082	1	0	60	1	60	486406
2001	122522	1	0	60	1	60	1207671
2002	102481	1	1	30	1	30	417918
2003	119005	1	1	30	1	30	248678
2004							572324

¹ HIP = data collected under the HIP program, SIS = years when seasons within a season (partial season) were adopted in the Mississippi Flyway, and MWS is the Midwinter Survey population estimate from the Mississippi flyway.

Appendix 11. Pacific flyway total pintail harvest and harvest regulations from 1979-2003.

Year	Harvest	HIP	SIS	Days	Bag	DaysBag	MWS
1979	829302	0	0	93	7	651	3265814
1980	633307	0	0	93	7	651	4015739
1981	403865	0	0	93	7	651	2508739
1982	467575	0	0	93	7	651	1831832
1983	465088	0	0	93	7	651	1181335
1984	312488	0	0	93	5	465	2411716
1985	292708	0	0	79	5	395	859305
1986	274953	0	0	79	4	316	1254794
1987	311406	0	0	79	4	316	663212
1988	116304	0	0	59	1	59	1262689
1989	139507	0	0	59	1	59	685403
1990	133154	0	0	59	1	59	888876
1991	126404	0	0	59	1	59	1051819
1992	116312	0	0	59	1	59	773548
1993	140895	0	0	59	1	59	741120
1994	150376	0	0	69	1	69	1055970
1995	258506	0	0	93	2	186	1012086
1996	280743	0	0	93	2	186	1435296
1997	338312	0	0	107	3	321	962026
1998	237276	0	0	107	1	107	1278494
1999	191994	0	0	107	1	107	1129553
2000	159242	0	0	107	1	107	1444171
2001	133307	0	0	107	1	107	1710149
1999	221850	1	0	107	1	107	1129553
2000	183950	1	0	107	1	107	1444171
2001	146169	1	0	107	1	107	1710149
2002	132151	1	1	60	1	60	1081216
2003	134936	1	1	60	1	60	1139062
2004							1219077

¹ HIP = data collected under the HIP program, SIS = years when seasons within a season (partial season) were adopted in the Pacific Flyway, and MWS is the Midwinter Survey population estimate from the Pacific flyway.

Appendix 12. Continental pintail harvest, 1979-2003.

Year	PF	CF	MF	AF	AK	Canada	AK/Can	Total
1979	829302	228947	213601	48462	16315	145609	161925	1482237
1980	633307	193244	215811	38869	23896	128762	152660	1233890
1981	403865	151023	207864	27891	18105	110980	129087	919728
1982	467575	158994	126568	38632	11755	104790	116546	908313
1983	465088	139077	187365	18636	13206	101784	114991	925156
1984	312488	165804	153680	34658	16465	103407	119873	786503
1985	292708	83914	124920	21685	13341	91099	104441	627667
1986	274953	72071	90350	19033	13061	59979	73041	529448
1987	311406	122420	88305	15788	10600	67172	77774	615690
1988	116304	36387	39225	7447	10509	69346	79858	279218
1989	139507	43594	65055	14588	11080	62947	74028	336771
1990	133154	43206	49487	10493	10284	71624	81909	318248
1991	126404	28687	40319	14201	6518	35212	41731	251342
1992	116312	32095	56520	12470	10613	33408	44022	261418
1993	140895	42274	52635	12923	4966	37741	41972	291434
1994	150376	61456	81147	18340	5087	44431	49782	360837
1995	258506	94840	136099	33163	8766	44299	52957	575673
1996	280743	96634	123817	19270	13641	52689	65007	586793
1997	338312	189211	144758	24010	10310	60750	71067	767350
1998	237276	124994	176990	33594	12984	59843	72828	645681
1999	191994	138397	167666	29527	10283	55683	--	--
2000	159242	136592	161476	22384	16856	58339	--	--
2001	133307	151917	130894	19950	9277	39643	--	--
1999	221850	133317	148299	25200	10854	55683	66537	595203
2000	183950	134252	155082	20752	17213	58339	75552	569589
2001	146169	134612	122522	19276	11306	39643	50949	473528
2002	132151	60407	102481	17089	11281	57028	68309	380437
2003	134936	55641	119005	18134	9084	47950	57034	384750
2004								

Notes: (1) Data above the line use U.S. harvest estimates from the MQS, data below the line use estimates from HIP. (2) In years when both MQS and HIP estimates were available, only the HIP estimates were used to calculate totals.